

# *Future global mortality from changes in air pollution attributable to climate change*

Article

Accepted Version

Silva, R. A., West, J. J., Lamarque, J.-F., Shindell, D. T., Collins, W. J. ORCID: <https://orcid.org/0000-0002-7419-0850>, Faluvegi, G., Folberth, G. A., Horowitz, L. W., Nagashima, T., Naik, V., Rumbold, S. T. ORCID: <https://orcid.org/0000-0001-8138-4541>, Sudo, K., Takemura, T., Bergman, D., Cameron-Smith, P., Doherty, R. M., Josse, B., MacKenzie, I. A., Stevenson, D. S. and Zeng, G. (2017) Future global mortality from changes in air pollution attributable to climate change. *Nature Climate Change*, 7 (9). pp. 647-651. ISSN 1758-678X doi: <https://doi.org/10.1038/nclimate3354> Available at <https://centaur.reading.ac.uk/71591/>

It is advisable to refer to the publisher's version if you intend to cite from the work. See [Guidance on citing](#).

To link to this article DOI: <http://dx.doi.org/10.1038/nclimate3354>

Publisher: Nature Publishing Group

All outputs in CentAUR are protected by Intellectual Property Rights law, including copyright law. Copyright and IPR is retained by the creators or other copyright holders. Terms and conditions for use of this material are defined in the [End User Agreement](#).

[www.reading.ac.uk/centaur](http://www.reading.ac.uk/centaur)

**CentAUR**

Central Archive at the University of Reading

Reading's research outputs online

1 **FUTURE GLOBAL MORTALITY FROM CHANGES IN AIR POLLUTION**  
2 **ATTRIBUTABLE TO CLIMATE CHANGE**

3 **Raquel A. Silva<sup>1,2</sup>, J. Jason West<sup>1,\*</sup>, Jean-François Lamarque<sup>3</sup>, Drew T. Shindell<sup>4</sup>,**  
4 **William J. Collins<sup>5</sup>, Greg Faluvegi<sup>6</sup>, Gerd A. Folberth<sup>7</sup>, Larry W. Horowitz<sup>8</sup>, Tatsuya**  
5 **Nagashima<sup>9</sup>, Vaishali Naik<sup>10</sup>, Steven T. Rumbold<sup>11</sup>, Kengo Sudo<sup>12</sup>, Toshihiko**  
6 **Takemura<sup>13</sup>, Daniel Bergmann<sup>14</sup>, Philip Cameron-Smith<sup>14</sup>, Ruth M. Doherty<sup>15</sup>,**  
7 **Beatrice Josse<sup>16</sup>, Ian A. MacKenzie<sup>15</sup>, David S. Stevenson<sup>15</sup>, and Guang Zeng<sup>17</sup>**

8  
9 <sup>1</sup> Environmental Sciences and Engineering, University of North Carolina, Chapel Hill,  
10 North Carolina

11 <sup>2</sup> Now: Oak Ridge Institute for Science and Education at US Environmental Protection  
12 Agency, Research Triangle Park, North Carolina

13 <sup>3</sup> NCAR Earth System Laboratory, National Center for Atmospheric Research, Boulder,  
14 Colorado

15 <sup>4</sup> Nicholas School of the Environment, Duke University, Durham, North Carolina

16 <sup>5</sup> Department of Meteorology, University of Reading, Reading, United Kingdom

17 <sup>6</sup> NASA Goddard Institute for Space Studies and Columbia Earth Institute, New York,  
18 New York

19 <sup>7</sup> Met Office Hadley Centre for Climate Prediction, Exeter, United Kingdom

20 <sup>8</sup> NOAA Geophysical Fluid Dynamics Laboratory, Princeton, New Jersey

21 <sup>9</sup> National Institute for Environmental Studies, Tsukuba, Japan

22 <sup>10</sup> UCAR/NOAA Geophysical Fluid Dynamics Laboratory, Princeton, New Jersey

23 <sup>11</sup> National Centre for Atmospheric Science, University of Reading, Reading, United  
24 Kingdom

25 <sup>12</sup> Earth and Environmental Science, Graduate School of Environmental Studies,  
26 Nagoya University, Nagoya, Japan

27 <sup>13</sup> Research Institute for Applied Mechanics, Kyushu University, Fukuoka, Japan

28 <sup>14</sup> Lawrence Livermore National Laboratory, Livermore, California

29 <sup>15</sup> School of GeoSciences, University of Edinburgh, Edinburgh, United Kingdom

30 <sup>16</sup> GAME/CNRM, Météo-France, CNRS—Centre National de Recherches  
31 Meteorologiques, Toulouse, France

32 <sup>17</sup> National Institute of Water and Atmospheric Research, Wellington, New Zealand

33  
34

35 Correspondence should be addressed to: J. J. West (jjwest@email.unc.edu)

36

37

38

39

40 Ground-level ozone and fine particulate matter (PM<sub>2.5</sub>) are associated with premature human  
41 mortality<sup>1-4</sup>; their future concentrations depend on changes in emissions, which dominate the  
42 near-term<sup>5</sup>, and on climate change<sup>6,7</sup>. Previous global studies of the air quality-related health  
43 effects of future climate change<sup>8,9</sup> used single atmospheric models. However, in related studies,  
44 mortality results differ among models<sup>10-12</sup>. Here we use an ensemble of global chemistry-climate  
45 models<sup>13</sup> to show that premature mortality from changes in air pollution attributable to climate  
46 change, under the high greenhouse gas scenario RCP8.5<sup>14</sup>, is likely positive. We estimate 3,340  
47 (-30,300 to 47,100) ozone-related deaths in 2030, relative to 2000 climate, and 43,600 (-195,000  
48 to 237,000) in 2100 (14% of the increase in global ozone-related mortality). For PM<sub>2.5</sub>, we  
49 estimate 55,600 (-34,300 to 164,000) deaths in 2030 and 215,000 (-76,100 to 595,000) in 2100  
50 (countering by 16% the global decrease in PM<sub>2.5</sub>-related mortality). Premature mortality  
51 attributable to climate change is estimated to be positive in all regions except Africa, and is  
52 greatest in India and East Asia. Most individual models yield increased mortality from climate  
53 change, but some yield decreases, suggesting caution in interpreting results from a single model.  
54 Climate change mitigation will likely reduce air pollution-related mortality.

55 Climate change can affect air quality through several pathways, including changes in the  
56 ventilation and dilution of air pollutants, photochemical reaction rates, removal processes,  
57 stratosphere–troposphere exchange of ozone, wildfires, and natural biogenic and lightning  
58 emissions<sup>6,7</sup>. Overall, changes in these processes are expected to increase ozone in polluted  
59 regions during the warm season, especially in urban areas and during pollution episodes, but  
60 decrease ozone in remote regions due to greater water vapour concentrations leading to greater  
61 ozone destruction. These effects are exacerbated by the greater decomposition of reservoir  
62 species such as PAN<sup>7</sup>. PM<sub>2.5</sub> will also be affected by climate change, but impacts vary in sign

63 among models and show regional variation related to differences in precipitation, wildfires,  
64 biogenic emissions, PM<sub>2.5</sub> composition, and other factors.

65 Previous studies have examined the impact of future climate change on human health via air  
66 quality globally<sup>8-9,15</sup>, in the US<sup>10, 16-20</sup>, and in Europe<sup>21</sup>. However, only two studies have  
67 previously used an ensemble of models to assess air pollution-related mortality attributable to  
68 climate change: one for the US<sup>10</sup>, and our previous global work with the same ensemble used  
69 here, but evaluating the effects of historical climate change prior to 2000<sup>11</sup>. Both studies found a  
70 large spread of mortality outcomes depending on the atmospheric model used. Silva et al.<sup>11</sup>  
71 found that the multi-model average suggested a small detrimental effect of climate change on  
72 global present-day air pollution-related mortality, but individual models yielded estimates of  
73 opposing sign.

74 The Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP) ensemble  
75 (Supplementary Table 1) simulated air quality in 2000, and in 2030, 2050 and 2100 for the four  
76 global Representative Concentration Pathway scenarios (RCPs)<sup>22</sup>. We previously estimated  
77 future air pollution premature mortality under all four RCP scenarios, estimating the net effect of  
78 both emissions changes and climate change<sup>12</sup>. Under RCP8.5, ozone concentrations increase in  
79 most locations in 2100 relative to 2000, due to increases in methane emissions and the effect of  
80 climate change<sup>7,23</sup>, but PM<sub>2.5</sub> decreases in 2100 due to a projected decrease in particulate and  
81 precursor emissions<sup>24</sup>. These changes in pollutant concentrations lead to 316,000 (95% C.I.: -  
82 187,000 to 1.38 million) ozone-related excess deaths yr<sup>-1</sup> and -1.31 (-2.04 to -0.17) million  
83 PM<sub>2.5</sub>-related (avoided) deaths yr<sup>-1</sup> in 2100<sup>12</sup>. Here we present results from additional ACCMIP  
84 simulations that were designed to isolate the influences of future climate change under RCP8.5,  
85 by simulating the projected climates of 2030 and 2100 (imposed by prescribing sea-surface

86 temperatures, sea ice cover, and greenhouse gas concentrations for radiation) together with air  
87 pollutant emissions from 2000. The effects of climate change are then isolated by a difference  
88 with historical 2000 simulations. Premature mortality attributable to RCP8.5 climate change is  
89 estimated following the methods of Silva et al.<sup>12</sup>, including projected population and baseline  
90 mortality rates (see Methods), such that mortality estimates here can be compared directly with  
91 overall changes in air pollution-related mortality in RCP8.5.

92 We estimate that global ozone mortality attributable to RCP8.5 climate change will be 3,340 (-  
93 30,300 to 47,100) deaths yr<sup>-1</sup> in 2030 and 43,600 (-195,000 to 237,000) deaths yr<sup>-1</sup> in 2100  
94 (Figures 1a and 2a). In 2100, ozone mortality increases in most regions, especially in highly  
95 populated and highly polluted areas, with marked spatial differences within regions that include  
96 both positive and negative mortality changes (Figure 3a, Supplementary Table 2, Supplementary  
97 Figures 1 and 2a). The effect on ozone mortality in 2100 is greatest in East Asia (45,600 deaths  
98 yr<sup>-1</sup>, 41 deaths yr<sup>-1</sup> per million people), India (16,000 deaths yr<sup>-1</sup>, 8 deaths yr<sup>-1</sup> per million people)  
99 and North America (9,830 deaths yr<sup>-1</sup>, 13 deaths yr<sup>-1</sup> per million people), but some areas within  
100 these and other regions show decreases in mortality. East Asia has high mortality effects per  
101 person in part because of its higher projected mortality rate from respiratory diseases. Climate  
102 change contributes 14% of the overall increase in ozone mortality estimated for RCP8.5 in 2100  
103 relative to 2000<sup>12</sup>. However, three of 8 models in 2030 and three of 9 in 2100 show global  
104 decreases in ozone mortality due to climate change. For each model, the uncertainty range does  
105 not include zero; only the spread of models causes the overall uncertainty to span zero.  
106 Uncertainty in modeled ozone concentrations contributes over 97% to the overall uncertainty in  
107 both 2030 and 2100, with the remainder from uncertainties in relative risk (RR). Results from a  
108 sensitivity analysis using present-day population and baseline mortality rates (Table 1) show

109 32% and 67% lower mortality estimates in 2030 and 2100, respectively, largely because the  
110 projected baseline mortality rates of chronic respiratory diseases increase through 2100. The  
111 models agree that ozone will increase due to climate change in some polluted regions, notably  
112 the northeast US as found in other studies<sup>6</sup> and decrease in the tropics over the oceans  
113 (Supplementary Figures 3 and 4a). These changes are consistent with those analysed by Schnell  
114 et al.<sup>25</sup> for 2100, using four of these same models, and were attributed to a greater efficiency of  
115 precursor emissions to generate surface ozone in polluted regions, along with reductions in the  
116 export of precursors to downwind regions.

117 The impact of climate change on PM<sub>2.5</sub> mortality is estimated to result in 55,600 (-34,300 to  
118 164,000) deaths yr<sup>-1</sup> in 2030 and 215,000 (-76,100 to 595,000) deaths yr<sup>-1</sup> in 2100 (Figures 1b  
119 and 2b). Mean estimates of PM<sub>2.5</sub> mortality increase in 2100 in all regions except Africa (-25,200  
120 deaths yr<sup>-1</sup>) (Figure 3b, Supplementary Table 3, Supplementary Figure 2b). The greatest  
121 increases in mortality in 2100 occur in India (80,200 deaths yr<sup>-1</sup>, 40 deaths yr<sup>-1</sup> per million  
122 people), Middle East (50,400 deaths yr<sup>-1</sup>, 45 deaths yr<sup>-1</sup> per million people) and East Asia  
123 (47,200 deaths yr<sup>-1</sup>, 43 deaths yr<sup>-1</sup> per million people), although the Former Soviet Union shows  
124 greater mortality per million people in 2100 (11,800 deaths yr<sup>-1</sup>, 57 deaths yr<sup>-1</sup> per million  
125 people). Similar to ozone mortality, there are substantial spatial differences within each region,  
126 including both increases and decreases in mortality. For PM<sub>2.5</sub>, a large decrease in mortality is  
127 projected in RCP8.5 relative to 2000 (when accounting for changes in both emissions and  
128 climate)<sup>12</sup>, but climate change alone increases mortality, partially counteracting the decrease  
129 associated with declining emissions in RCP8.5. Without climate change, the decrease in PM<sub>2.5</sub>-  
130 related mortality would be roughly 16% greater in 2100 relative to 2000. Propagating  
131 uncertainty in RR to the mortality estimates leads to coefficients of variation (CVs) of 8-31%

132 (2030) and 11-46% (2100) for the different models, but the spread of model results increases  
133 overall CVs to 123% in 2030 and 106% in 2100. In both years, one model (GISS-E2-R) yields a  
134 decrease in global mortality from climate change while the other three (2030) or four (2100)  
135 show an increase. Uncertainty in modeled  $PM_{2.5}$  concentrations in 2000 makes a similar  
136 contribution to the overall uncertainty (50% in 2030 and 52% in 2100) compared with  
137 uncertainty in modeled  $PM_{2.5}$  concentrations in future years (50% in 2030, 48% in 2100).  
138 Uncertainty in RR makes a negligible contribution in both periods (<1%), as the multi-model  
139 mean is small and different models disagree on the sign of the influence. Considering present-  
140 day population and baseline mortality rates (Table 1), we estimate 23% and 33% lower mortality  
141 in 2030 and 2100, respectively, mostly associated with the increase in projected baseline  
142 mortality rates through 2100.

143  $PM_{2.5}$ -related mortality was estimated above for the sum of  $PM_{2.5}$  species reported by five  
144 models, using a common formula (see Methods), to increase the number of models considered  
145 and to increase consistency among  $PM_{2.5}$  estimates. Additionally, we present a sensitivity  
146 analysis considering the  $PM_{2.5}$  concentrations reported by four models using their own  $PM_{2.5}$   
147 formulas, for which multi-model average mortality results are modestly higher: 15% greater in  
148 2030 and 12% in 2100 (Supplementary Figure 5). The degree of agreement between the two  
149 estimates varies among the four models, and for one model (GISS-E2-R) the two sources of  
150  $PM_{2.5}$  estimates yield impacts of different sign in 2030.

151 There is considerable agreement among models regarding the increase in  $PM_{2.5}$  concentrations in  
152 many locations in 2100, including most polluted regions, due to RCP8.5 climate change  
153 (Supplementary Figure 4b). Allen et al.<sup>26</sup> analysed four of these same models in 2100 and found  
154 that global average surface  $PM_{2.5}$  concentrations increased due to climate change, reflecting



155 increases in nearly all relevant species for each model. They attributed this increase in  $PM_{2.5}$   
156 mainly to a decrease in wet deposition associated with less large-scale precipitation over land.  
157 Our multi-model mean estimates of global population-weighted changes for  $PM_{2.5}$  and individual  
158 species (Supplementary Table 4; Supplementary Figure 6) are similar to those of Allen et al.<sup>26</sup>.  
159 Unlike Allen et al.<sup>26</sup>, however, GISS-E2-R shows a net decrease in global population-weighted  
160 concentrations of total  $PM_{2.5}$  and of each  $PM_{2.5}$  species except sea salt, in 2100, likely due to  
161 projected concentration decreases over densely-populated eastern China. Models also differ  
162 strongly in the sign and magnitude of changes in dust, particularly over North Africa and the  
163 Middle East; HadGEM2 projects increases in  $PM_{2.5}$  for all species except dust, but a strong  
164 decrease in dust over the Middle East and South Asia. In Africa, the decrease in  $PM_{2.5}$  near the  
165 equator is likely caused by increased precipitation, whereas  $PM_{2.5}$  increases are associated with  
166 precipitation decreases in Southern Africa<sup>26</sup>. Differences in  $PM_{2.5}$  (and ozone) responses to  
167 climate change among models likely result from differences in large-scale meteorological  
168 changes, and different treatments of atmospheric chemistry and feedback processes among the  
169 models (such as the response of dust to climate change).

170 In the US, our multi-model mean mortality estimates for the impact of RCP8.5 climate change  
171 for ozone (1,130 deaths  $yr^{-1}$  in 2030; 8,810 deaths  $yr^{-1}$  in 2100) compare well with those of Fann  
172 et al.<sup>20</sup>, who report 420 to 1900 ozone-related deaths  $yr^{-1}$  for RCP8.5 climate change in 2030,  
173 despite differences in concentration-response functions and population and baseline mortality  
174 projections. These results for ozone and those for  $PM_{2.5}$  (6,900 deaths  $yr^{-1}$  in 2030; 19,400 deaths  
175  $yr^{-1}$  in 2100) are also consistent with the increases in mortality and spatial heterogeneity  
176 attributed to climate change in 2050 by Bell et al.<sup>16</sup> for ozone and Tagaris et al.<sup>17</sup> for ozone and  
177  $PM_{2.5}$ , although these studies used different climate change scenarios besides other

178 methodological differences. Across models, our estimates for ozone mortality in the US vary  
179 between -435 and 4,750 deaths yr<sup>-1</sup> in 2030 and between -1,820 and 27,012 deaths yr<sup>-1</sup> in 2100.  
180 This spread of model results, with a few models suggesting avoided mortality due to climate  
181 change, is similar to that of Post et al.<sup>10</sup> (-600 to 2,500 deaths yr<sup>-1</sup> in 2050) using SRES scenarios  
182 of GHG emissions. Similarly, results show spatial heterogeneity within several regions (Figure  
183 2) that is similar to Post et al.<sup>10</sup> for the US and Orru et al.<sup>21</sup> for Europe.

184 The spread of results among models highlights the uncertainty in the effect of climate change on  
185 air quality. Further improvements in chemistry climate models are needed to better model the  
186 interaction and feedbacks between climate and air quality, including the sensitivity of biogenic  
187 emissions to climate change, the effects of meteorological changes on air quality (e.g., aerosol-  
188 cloud interactions, secondary aerosol formation, wet deposition, and gas-aerosol partitioning),  
189 and the impact of climate change on wildfires. Stratosphere-troposphere exchange of ozone is  
190 also important, as is the impact of land use changes on regional climate and air pollution. Our  
191 results are specific to climate change as projected under RCP8.5 and would differ for other  
192 scenarios. We estimate the effect of climate change as the difference between simulations with  
193 future climate and year 2000 climate, both with year 2000 emissions, although global emissions  
194 of PM<sub>2.5</sub> and its main precursors decrease under RCP8.5. Had we instead modelled future  
195 emissions with present vs. future climate, we would likely have attributed smaller changes in air  
196 pollution and mortality to climate change, given the projected emission reductions. Whereas the  
197 net effect of missing and uncertain processes does not clearly indicate an under- or overestimate  
198 for the effect of climate change on air quality, we likely underestimate the magnitude of the  
199 health impact by omitting mortality for people under 25, and morbidity effects. We also neglect  
200 possible synergistic effects of a warmer climate to modify air pollution-mortality relationships.

201 Although a few studies have suggested stronger relationships between ozone<sup>27</sup> and PM<sub>2.5</sub><sup>28</sup> and  
202 health at higher temperatures, there is insufficient evidence to include those effects here.

203 Despite these uncertainties, this study is the first to use a multi-model ensemble to show that  
204 global air pollution-related mortality attributable to climate change is likely positive. The spread  
205 of results among models within the ensemble, including differences in the sign of global and  
206 regional mortality estimates, suggests that results from studies using a single model and a small  
207 number of model years should be interpreted cautiously. Actions to mitigate climate change,  
208 such as reductions in long-lived GHG emissions, will likely benefit human health by reducing  
209 the effect of climate change on air quality in many locations. These health benefits are likely to  
210 be smaller than those from reducing co-emitted air pollutants<sup>29</sup>, but both types of health benefits  
211 via changes in air quality would add to reductions in many other influences of climate change on  
212 human health<sup>30</sup>.

213

## 214 **Additional information**

215 Supplementary information is available in the online version of the paper.

## 216 **References**

- 217 1. Jerrett, M. *et al.* Long-Term Ozone Exposure and Mortality. *N. Engl. J. Med.* **360**, 1085–  
218 1095 (2009).
- 219 2. Krewski, D. *et al.* Extended Follow-Up and Spatial Analysis of the American Cancer  
220 Society Study Linking Particulate Air Pollution and Mortality. *Respir. Rep. Heal. Eff.*  
221 *Inst.* **140**, 5–114 (2009).
- 222 3. Lepeule, J., Laden, F., Dockery, D. & Schwartz, J. Chronic Exposure to Fine Particles  
223 and Mortality: An Extended Follow-up of the Harvard Six Cities Study from 1974 to  
224 2009. *Environ. Health Perspect.* **120**, 965–970 (2012).

- 225 4. Burnett, R. T. *et al.* An Integrated Risk Function for Estimating the Global Burden of  
226 Disease Attributable to Ambient Fine Particulate Matter Exposure. *Environ. Health*  
227 *Perspect.* **122**, 397–403 (2014).
- 228 5. Kirtman, B. *et al.* Near-term Climate Change: Projections and Predictability. In: *Climate*  
229 *Change 2013: The Physical Science Basis (Chapter 11)*. Contribution of Working Group  
230 I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change  
231 [Stocker, T.F.*et al.* (eds.)]. Cambridge University Press, Cambridge, United Kingdom  
232 and New York, NY, USA (2013).
- 233 6. Fiore, A. M., Naik, V. & Leibensperger, E. M. Air Quality and Climate Connections. *J*  
234 *Air Waste Manage. Assoc.* **65**, 645–685 (2015).
- 235 7. von Schneidmesser, E. *et al.* Chemistry and the Linkages between Air Quality and  
236 Climate Change. *Chem. Rev.* 150430065937004 (2015).  
237 doi:10.1021/acs.chemrev.5b00089
- 238 8. West, J. J., Szopa, S. & Hauglustaine, D. A. Human mortality effects of future  
239 concentrations of tropospheric ozone. *C. R. Geosci.* **339**, 775–83, 2007.
- 240 9. Selin, N. E. *et al.* Global health and economic impacts of future ozone pollution. *Environ.*  
241 *Res. Lett.* **4**, 044014 (2009).
- 242 10. Post, E. S. *et al.* Variation in Estimated Ozone-Related Health Impacts of Climate  
243 Change due to Modeling Choices and Assumptions. *Environ. Health Perspect.* **120**,  
244 1559–1564 (2012).
- 245 11. Silva, R. A. *et al.* Global premature mortality due to anthropogenic outdoor air pollution  
246 and the contribution of past climate change. *Environ. Res. Lett.* **8**, 034005 (2013).
- 247 12. Silva, R. A. *et al.* The effect of future ambient air pollution on human premature  
248 mortality to 2100 using output from the ACCMIP model ensemble. *Atmos. Chem. Phys.*  
249 **16**, 9847–9862 (2016).
- 250 13. Lamarque, J. F. *et al.* The Atmospheric Chemistry and Climate Model Intercomparison  
251 Project (ACCMIP): Overview and description of models, simulations and climate  
252 diagnostics. *Geosci. Model Dev.* **6**, 179–206 (2013).
- 253 14. Stevenson, D. S. *et al.* Tropospheric ozone changes, radiative forcing and attribution to  
254 emissions in the Atmospheric Chemistry and Climate Model Intercomparison Project  
255 (ACCMIP). *Atmos. Chem. Phys.* **13**, 3063–3085 (2013).
- 256 15. Fang, Y., Mauzerall, D. L., Liu, J., Fiore, A. M. & Horowitz, L. W. Impacts of 21st  
257 century climate change on global air pollution-related premature mortality. *Clim. Change*  
258 **121**(2), 239–253 (2013)
- 259 16. Bell, M. L. *et al.* Climate change, ambient ozone, and health in 50 US cities. *Clim.*  
260 *Change* **82**, 61–76 (2007).
- 261 17. Tagaris, E. *et al.* Potential Impact of Climate Change on Air Pollution-Related Human  
262 Health Effects. *Environ. Sci. Technol.* **43**, 4979–4988 (2009).
- 263 18. Chang, H. H., Zhou, J. & Fuentes, M. Impact of Climate Change on Ambient Ozone  
264 Level and Mortality in Southeastern United States. *Int. J. Environ. Res. Public Health* **7**,  
265 2866–2880 (2010).
- 266 19. Sheffield, P. E., Knowlton, K., Carr, J. L. & Kinney, P. L. Modeling of Regional Climate  
267 Change Effects on Ground-Level Ozone and Childhood Asthma. *Am. J. Prev. Med.* **41**,  
268 251–257 (2011).

- 269 20. Fann, N. *et al.* The geographic distribution and economic value of climate change-related  
270 ozone health impacts in the United States in 2030. *J. Air Waste Manage. Assoc.* **65**, 570–  
271 580 (2015).
- 272 21. Orru, H. *et al.* Impact of climate change on ozone-related mortality and morbidity in  
273 Europe. *Eur. Respir. J.* **41**, 285–294 (2013).
- 274 22. van Vuuren, D. P. *et al.* The representative concentration pathways: an overview. *Clim.*  
275 *Change* **109**, 5–31 (2011).
- 276 23. Young, P. J. *et al.* Pre-industrial to end 21st century projections of tropospheric ozone  
277 from the Atmospheric Chemistry and Climate Model Intercomparison Project  
278 (ACCMIP). *Atmos. Chem. Phys.* **13**, 2063–2090 (2013).
- 279 24. Shindell, D. T. *et al.* Radiative forcing in the ACCMIP historical and future climate  
280 simulations. *Atmos. Chem. Phys.* **13**, 2939–2974 (2013).
- 281 25. Schnell, J. L. *et al.* Effect of climate change on surface ozone over North America,  
282 Europe, and East Asia. *Geophys. Res. Lett.*, **43**, 3509–3518 (2016).
- 283 26. Allen, R.J., Landuyt, W. & Rumbold, T. An increase in aerosol burden and radiative  
284 effects in a warmer world. *Nat. Clim. Change* **6**, 269–274 (2016).
- 285 27. Wilson, A., Rappold, A. G., Neas, L. M. & Reich, B. J. Modeling the effect of  
286 temperature on ozone-related mortality. *Ann. Appl. Stat.* **8**, 1728–1749 (2014).
- 287 28. Ren, C., Williams, G. M. & Tong, S. Does Particulate Matter Modify the Association  
288 between Temperature and Cardiorespiratory Diseases?. *Environ. Health Perspect.* **114**,  
289 1690–1696 (2006).
- 290 29. West, J. J. *et al.* Co-benefits of Global Greenhouse Gas Mitigation for Future Air Quality  
291 and Human Health. *Nat. Clim. Change* **3**, 885–889 (2013).
- 292 30. Smith, K.R. *et al.* Human health: impacts, adaptation, and co-benefits. In: *Climate*  
293 *Change 2014: Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral*  
294 *Aspects*. Contribution of Working Group II to the Fifth Assessment Report of the  
295 Intergovernmental Panel on Climate Change [Field, C.B. *et al.* (eds.)], pp. 709-754  
296 (2014).

297  
298

## 299 **Acknowledgements**

300 This research was funded by NIEHS grant #1 R21 ES022600-01, a fellowship from the  
301 Portuguese Foundation for Science and Technology, and by a Dissertation Completion  
302 Fellowship from The Graduate School (UNC – Chapel Hill). We thank Karin Yeatts (Gillings  
303 School of Global Public Health, UNC – Chapel Hill), Colin Mathers (WHO), Peter Speyer  
304 (IHME), and Amanda Henley (Davis Library Research & Instructional Services, UNC – Chapel  
305 Hill). The work of DB and PC was funded by the U.S. Dept. of Energy (BER), performed under  
306 the auspices of LLNL under Contract DE-AC52-07NA27344, and used the supercomputing

307 resources of NERSC under contract No. DE-AC02-05CH11231. RD, IM and DS acknowledge  
308 ARCHER supercomputing resources and funding under the UK Natural Environment Research  
309 Council grant: NE/I008063/1. GZ acknowledges the NZ eScience Infrastructure which is funded  
310 jointly by NeSI's collaborator institutions and through the MBIE's Research Infrastructure  
311 programme. GAF has received funding from BEIS under the Hadley Centre Climate Programme  
312 contract (GA01101) and from the European Union's Horizon 2020 research and innovation  
313 programme under grant agreement No 641816 (CRESCENDO). DTS and GF acknowledge the  
314 NASA High-End Computing Program through the NASA Center for Climate Simulation at  
315 Goddard Space Flight Center for computational resources.

316

317 **Author contributions:** JJW, JFL, DTS and RAS conceived the study. All other co-authors  
318 conducted the model simulations. RAS processed model output and estimated human mortality.  
319 RAS and JJW analyzed results. RAS and JJW prepared the manuscript and all co-authors  
320 commented on it.

321 **Competing Financial Interests:** All authors declare that they do not have any competing  
322 financial interests.

323

324

325 **Figure Legends:**

326 Figure 1 – Impact of RCP8.5 climate change on global mortality for individual models and the  
327 multi-model average. Estimates are for 2030 and 2100 for (a) ozone respiratory mortality (9  
328 models) and (b) PM2.5 IHD+STROKE+COPD+LC mortality (5 models). PM2.5 is calculated as  
329 a sum of species. Uncertainty for each model is the 95% CI taking into account uncertainty in  
330 RR. Uncertainty for the multi-model average is the 95% CI including uncertainty in RR and  
331 across models.

332  
333 Figure 2 – Geographical impact of climate change on mortality. Estimates are for 2030 and 2100  
334 for (a) ozone respiratory mortality and (b) PM2.5 IHD+STROKE+COPD+LC mortality,  
335 showing the multi-model average in each 0.5°x0.5° grid cell. PM2.5 is calculated as a sum of  
336 species.

337  
338 Figure 3 – Projected mortality for ten world regions. Estimates are for 2030 and 2100 for (a)  
339 ozone respiratory mortality and (b) PM2.5 IHD+STROKE+COPD+LC mortality, showing the  
340 multi-model regional average. PM2.5 is calculated as a sum of species. Uncertainty for the multi-  
341 model regional average is the 95% CI including uncertainty in RR and across models. World  
342 regions are shown in Supplementary Figure 1.

343

344

345 Table 1 – Sensitivity analysis for changes in global air pollution-related mortality attributable to  
346 climate change. Estimates are for multi-model averages (deaths yr<sup>-1</sup>) for the deterministic  
347 results.

348

	PM <sub>2.5</sub> -related mortality		Ozone-related mortality	
	2030	2100	2030	2100
Base results	56,300	218,000	10,700	128,000
PM <sub>2.5</sub> using Krewski et al. <sup>2</sup>	66,200	318,000	--	--
Present-day (2011) population	35,500	93,800	2,970	59,400
Present-day (2010) baseline mortality rates	69,600	510,000	2,790	13,300
Present-day population and baseline mortality rates	43,300	144,000	2,300	14,500

349

350

351

352



## 353 **Methods**

354 The Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP)<sup>13</sup> included  
355 contributions from 14 modelling groups, of which 9 completed simulations that are used here  
356 (Supplementary Table 1). ACCMIP models incorporate chemistry-climate interactions, including  
357 mechanisms by which climate change affects ozone and PM<sub>2.5</sub>, although models do not all  
358 include the same interactions, and do not always agree on their net effects<sup>7</sup>. Of these nine, three  
359 models are not truly coupled chemistry-climate models: MOCAGE is a chemical transport model  
360 driven by external meteorology, and UM-CAM and STOC-HadAM3 do not model the feedback  
361 of chemistry on climate<sup>13</sup>. As a result, these models do not fully capture the effects of changes in  
362 air pollutant concentrations on processes that affect meteorology, such as through radiative  
363 transfer and clouds. Prescribed anthropogenic and biomass burning emissions were very similar  
364 for the different models, but they used different natural emissions (e.g. biogenic volatile organic  
365 compounds, ocean emissions, soil and lightning NO<sub>x</sub>)<sup>14,23</sup>. Modelled 2000 concentrations show  
366 good agreement with observations for ozone<sup>23</sup> and PM<sub>2.5</sub><sup>24</sup>, although models tend to overestimate  
367 ozone in the Northern Hemisphere and underestimate it in the Southern Hemisphere, and to  
368 underestimate PM<sub>2.5</sub>, particularly in East Asia.

369 We isolate the effect of climate change on air quality as the difference in concentrations between  
370 ACCMIP simulations using year 2000 emissions together with future year climate, imposed by  
371 prescribing RCP8.5<sup>31</sup> sea surface temperatures, sea ice cover, and GHGs (for radiation) for 2030  
372 and 2100 (referred to as “Em2000Cl2030” and “Em2000Cl2100”), and simulations with 2000  
373 emissions and climate (“acchist2000”)<sup>13</sup>. We analyse results from the nine models reporting  
374 ozone from the Em2000Cl2030/2100 simulations, and the five reporting PM<sub>2.5</sub> (Supplementary  
375 Table 1). Ozone and PM<sub>2.5</sub> species surface concentrations from each model are calculated in each

376 grid cell, after regridding output from the native horizontal resolutions of each model (1.9°x1.2°  
377 to 5°x5°) to a common 0.5°x0.5° resolution. To be consistent with the epidemiological studies  
378 considered<sup>1,4</sup>, we use the seasonal average of daily 1-hr maximum ozone concentrations for the  
379 six consecutive months with highest concentrations in each grid cell, and annual average PM<sub>2.5</sub>  
380 concentration.

381 Seven of the nine models with Em2000Cl2030/2100 simulations reported both hourly and  
382 monthly ozone concentrations, while two reported only monthly values. We calculate the ratio  
383 of the 6-month average of daily 1-hr maximum concentrations to the annual average  
384 concentrations, for each grid cell and each year, for those models that reported both hourly and  
385 monthly concentrations; then, we apply that ratio to the annual average ozone concentrations for  
386 the other two models, following Silva *et al.*<sup>11,12</sup>.

387 We calculate PM<sub>2.5</sub> concentration using the sum of PM<sub>2.5</sub> species mass mixing ratios reported by  
388 five models and a common formula:

$$389 \text{ PM}_{2.5} = \text{BC} + \text{OA} + \text{SO}_4 + \text{SOA} + \text{NH}_4 + 0.25 * \text{SS} + 0.1 * \text{Dust},$$

390 where BC – Black Carbon, OA – (Primary) Organic Aerosol corrected to include species other  
391 than carbon, NH<sub>4</sub> – NH<sub>4</sub> in ammonium sulfate, SOA – Secondary Organic Aerosol, and SS –  
392 Sea Salt, as had been done previously by Fiore *et al.*<sup>33</sup> and Silva *et al.*<sup>11,12</sup>. The factors 0.25 and  
393 0.1 are intended to approximate the fractions of sea salt and dust that are in the PM<sub>2.5</sub> size range.  
394 Nitrate was reported by three models, but we chose to omit nitrate from our PM<sub>2.5</sub> formula to  
395 avoid imposing changes inconsistent with the effect of climate change for other models,  
396 following Silva *et al.*<sup>11</sup>, although nitrate was included in estimates of total PM<sub>2.5</sub> by Silva *et al.*<sup>12</sup>.  
397 Four of these models also reported their own estimate of PM<sub>2.5</sub> (Supplementary Table 1).

398 The impact of climate change on global population-weighted differences (Em2000CI2030/2100  
399 minus acchist2000) in PM<sub>2.5</sub> and ozone concentrations for the different models are shown in  
400 Supplementary Tables 4 and 5, respectively, while regional multi-model average differences are  
401 shown in Supplementary Figures 7 and 8.

402 We estimate premature mortality by calculating the fraction of cause-specific mortality  
403 attributable to long-term changes in pollutant concentrations, using methods that are identical to  
404 those of Silva *et al.*<sup>12</sup>, so that mortality attributable to climate change can be compared simply  
405 with changes in mortality under the RCP scenarios. We use relative risks (RRs) from Jerrett *et*  
406 *al.*<sup>1</sup> for ozone and respiratory diseases and Burnett *et al.*<sup>4</sup> for PM<sub>2.5</sub> and cardiopulmonary diseases  
407 and lung cancer. Then, we apply that attributable fraction in each grid cell to future adult  
408 population (age 25 and older) and baseline mortality rates based on projections from the  
409 International Futures (IFs) integrated modelling system<sup>32</sup>. Using country-level projections per  
410 age group, we mapped and gridded to the 0.5°x0.5° grid assuming that the present-day spatial  
411 distribution of total population within each country is unchanged in the future, as well as the  
412 present-day ratio of baseline mortality for the specific causes included in the epidemiological  
413 studies and for three disease groups projected in IFs (chronic respiratory diseases, cardiovascular  
414 diseases and malignant neoplasms). We select population projections from IFs instead of those  
415 underlying RCP8.5 to ensure consistency between projections of population and baseline  
416 mortality, since the latter are not available for RCP8.5, and for consistency with Silva *et al.*<sup>12</sup>. IFs  
417 projections of future total population are lower than those of RCP8.5 (-5% in 2030 and -27% in  
418 2100) (Supplementary Figure 9). Had we used projections of population underlying RCP8.5, we  
419 would have likely estimated greater changes in premature mortality relative to 2000. IFs  
420 projections of baseline mortality rates reflect an aging population and regional demographic

421 changes, showing a steep rise in chronic respiratory diseases (roughly tripling globally by 2100),  
422 particularly in East Asia and India, some regional increases in cardiovascular diseases (e.g.  
423 Middle East, Africa), and global decreases in lung cancer.

424 Overall uncertainty in mortality estimates includes uncertainty from the RRs and from air  
425 pollutant concentrations. First, we conduct 1000 Monte Carlo (MC) simulations separately for  
426 each model-year to propagate uncertainty from the RRs to mortality estimates. For ozone, we use  
427 the 95% Confidence Intervals (CIs) for RR reported by Jerrett *et al.*<sup>1</sup> and assume a normal  
428 distribution, while for PM<sub>2.5</sub> we use the parameter values of Burnett *et al.*<sup>4</sup> for 1000 MC  
429 simulations. Then, we calculate the average and 95% CI for the pooled results of the 1000 MC  
430 simulations for each model to quantify the spread of model results. We do not include  
431 uncertainties associated with population and baseline mortality rates, since these are not reported.  
432 As ACCMIP models used the same anthropogenic and biomass burning emissions, we do not  
433 consider uncertainty in emissions inventories, however we acknowledge that this is an important  
434 source of uncertainty, especially in particular regions<sup>34-37</sup>. Our mortality estimates are affected by  
435 our choices of and underlying assumptions regarding concentration-response functions,  
436 population, and baseline mortality rates. Although a number of factors, such as vulnerability of  
437 the exposed population and PM<sub>2.5</sub> composition, vary spatially and possibly temporally, we  
438 assume that the RRs estimated for the present day apply on a global scale and in future time  
439 periods. Also, our assumption that the spatial distribution of population within each country is  
440 constant in the future likely understates the effects of rural-to-urban migration, which is currently  
441 underway and expected to continue. However, the effects of climate change on air pollutant  
442 concentrations may be somewhat spatially uniform (as opposed to changes in emissions), and the

443 coarse grid resolution of global models would not resolve air pollutant concentrations well in  
444 urban areas.

445

#### 446 **Data Availability**

447 Data used in this project are archived here:

448 Air pollutant concentrations: Atmospheric Chemistry & Climate Model Intercomparison Project  
449 (ACCMIP) datasets - <http://catalogue.ceda.ac.uk/uuid/b46c58786d3e5a3f985043166aeb862d> .

450 Data retrieved from 08/2012 to 12/2013.

451 Present-day population: Oak Ridge National Laboratory (ONRL) - LandScan 2011 Global  
452 Population Dataset, <http://spruce.lib.unc.edu.libproxy.lib.unc.edu/content/gis/LandScan/> . Data  
453 retrieved on 12/05/2012.

454 Present-day baseline mortality: Institute for Health Metrics and Evaluation (IHME): Global  
455 Burden of Disease Study 2010 (GBD 2010) Results by Cause 1990-2010 - Country Level,  
456 Seattle, United States, 2013.  
457 [https://cloud.ihme.washington.edu/index.php/s/d559026958b38c3f4d12029b36d783da?path=%2](https://cloud.ihme.washington.edu/index.php/s/d559026958b38c3f4d12029b36d783da?path=%2F2010)  
458 [F2010](https://cloud.ihme.washington.edu/index.php/s/d559026958b38c3f4d12029b36d783da?path=%2F2010) . Data retrieved from 12/2013 to 03/2014.

459 Future population and baseline mortality: Web-Based IFs - The International Futures (IFs)  
460 modeling system, version 6.54., [www.ifs.du.edu](http://www.ifs.du.edu) . Data retrieved on 07/2012.

461 IER model: Global Burden of Disease Study 2010. Global Burden of Disease Study 2010 (GBD  
462 2010) - Ambient Air Pollution Risk Model 1990 - 2010. Seattle, United States: Institute for

463 Health Metrics and Evaluation (IHME), 2013. <http://ghdx.healthdata.org/record/global-burden->  
464 [disease-study-2010-gbd-2010-ambient-air-pollution-risk-model-1990-2010](http://ghdx.healthdata.org/record/global-burden-disease-study-2010-gbd-2010-ambient-air-pollution-risk-model-1990-2010) . Data retrieved on  
465 11/08/2013.

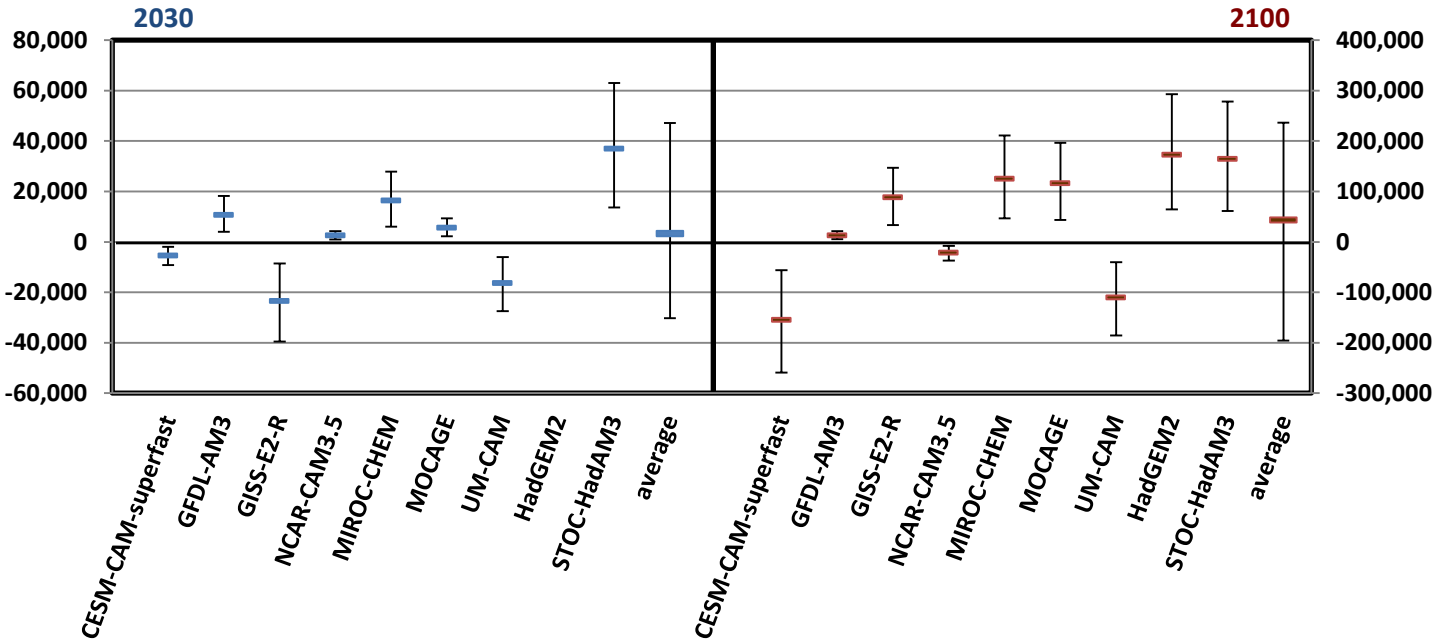
## 466 **References**

- 467 31. Riahi, K. *et al.* RCP 8.5 - A scenario of comparatively high greenhouse gas emissions.  
468 *Clim. Change* **109**, 33–57 (2011).
- 469 32. Hughes, B. B. *et al.* Projections of global health outcomes from 2005 to 2060 using the  
470 International Futures integrated forecasting model. *Bull. World Health Organ.* **89**, 478–486  
471 (2011).
- 472 33. Fiore, A. M. *et al.* Global air quality and climate. *Chem. Soc. Rev.* **41**, 6663–6683 (2012).
- 473 34. Bond, T. C. *et al.* Historical emissions of black and organic carbon aerosol from energy-  
474 related combustion, 1850-2000. *Glob. Biogeochem. Cycles* **21**, GB2018 (2007).
- 475 35. Schopp, W., Klimont, Z., Suutari, R. & Cofala, J. Uncertainty analysis of emissions  
476 estimates in the RAINS integrated assessment model. *Environ. Sci. Policy* **8**, 601-613  
477 (2005).
- 478 36. Smith, S. J. *et al.* Anthropogenic sulfur dioxide emissions: 1850-2005. *Atmos. Chem. Phys.*  
479 **11**, 11011-11116 (2011).
- 480 37. Granier, C. *et al.* Evolution of anthropogenic and biomass burning emissions of air  
481 pollutants at global and regional scales during the 1980-2010 period, *Clim. Change*, **109**,  
482 163-190 (2011).

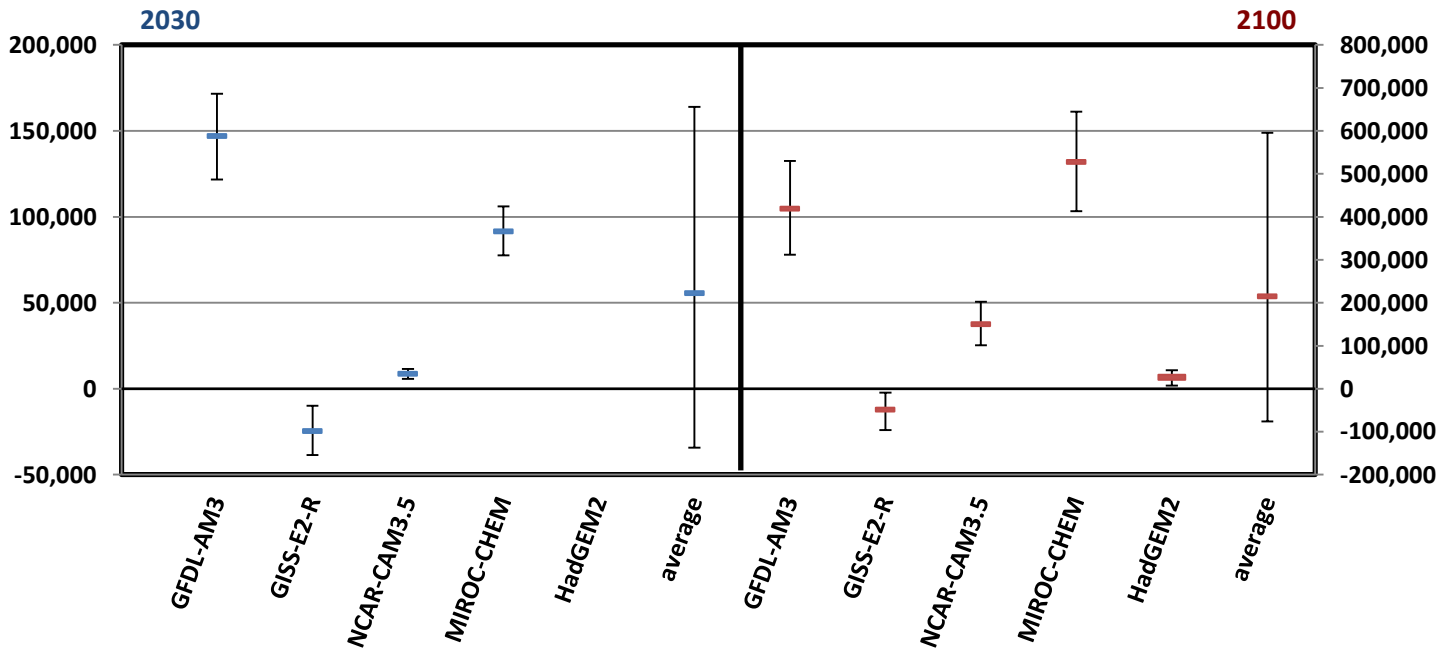
484

485

**a. Ozone mortality**  
(deaths yr<sup>-1</sup>)

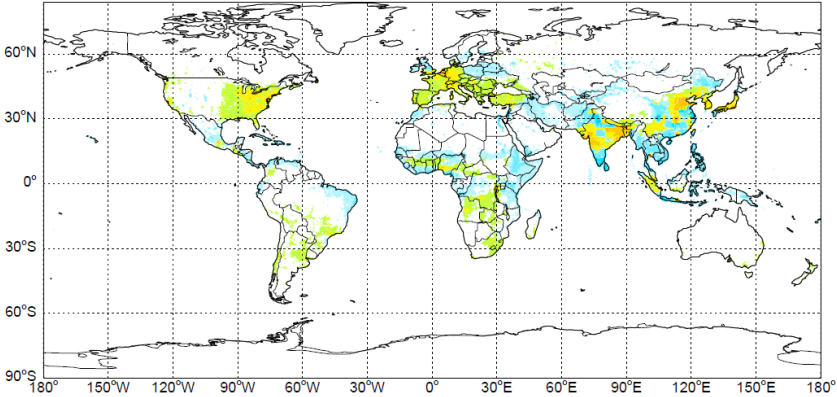


**b. PM<sub>2.5</sub> mortality**  
(deaths yr<sup>-1</sup>)

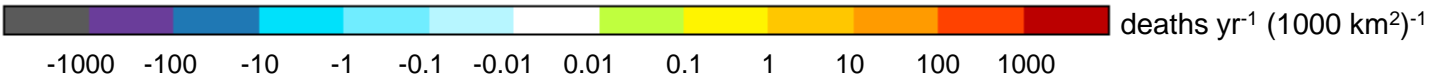
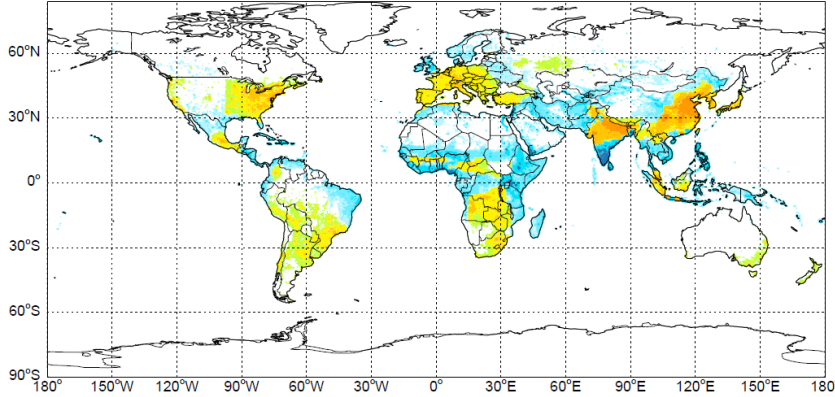


**a. Ozone mortality**

**2030** **8 models**

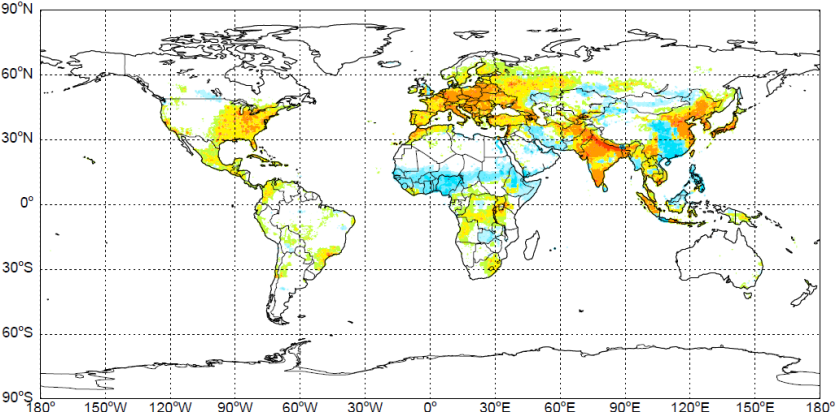


**2100** **9 models**

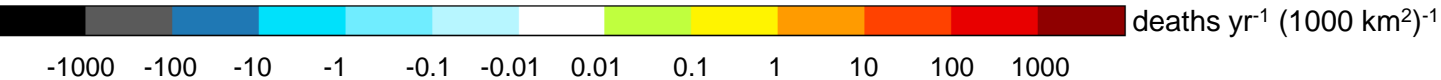
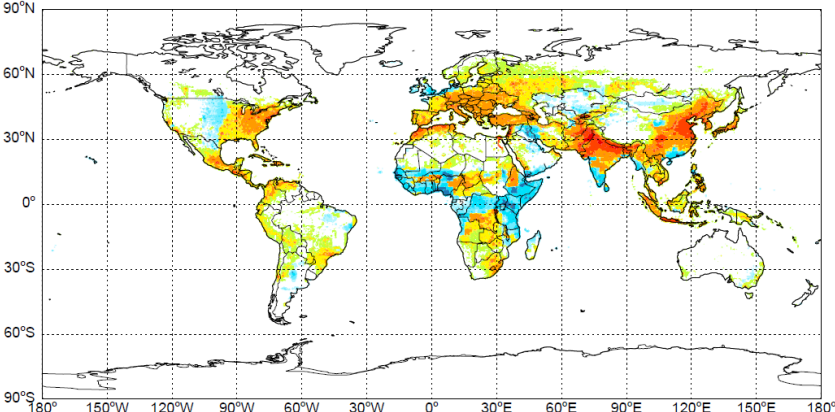


**b. PM<sub>2.5</sub> mortality**

**2030** **4 models**

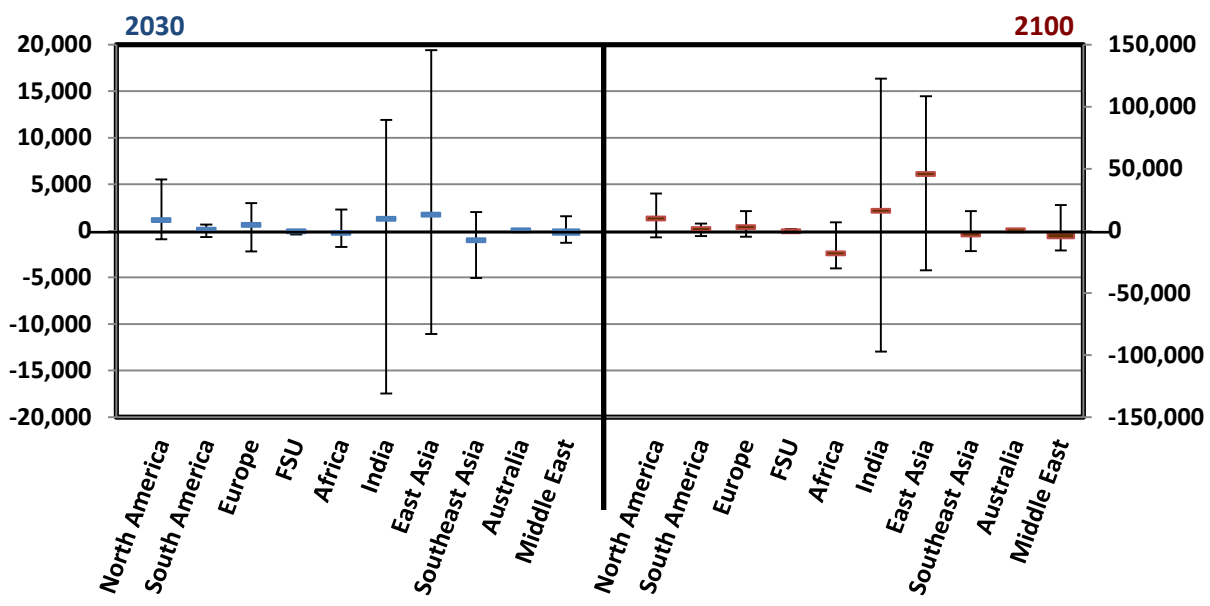


**2100** **5 models**





**a. Ozone mortality (deaths yr<sup>-1</sup>)**



**b. PM<sub>2.5</sub> mortality (deaths yr<sup>-1</sup>)**

