

# *Impact of plant choice on rainfall runoff delay and reduction by hedge species*

Article

Accepted Version

Blanusa, T. and Hadley, J. (2019) Impact of plant choice on rainfall runoff delay and reduction by hedge species. *Landscape and Ecological Engineering*, 15 (4). pp. 401-411. ISSN 1860-188X doi: <https://doi.org/10.1007/s11355-019-00390-x> Available at <https://centaur.reading.ac.uk/85330/>

It is advisable to refer to the publisher's version if you intend to cite from the work. See [Guidance on citing](#).

To link to this article DOI: <http://dx.doi.org/10.1007/s11355-019-00390-x>

Publisher: Springer

All outputs in CentAUR are protected by Intellectual Property Rights law, including copyright law. Copyright and IPR is retained by the creators or other copyright holders. Terms and conditions for use of this material are defined in the [End User Agreement](#).

[www.reading.ac.uk/centaur](http://www.reading.ac.uk/centaur)

**CentAUR**

Central Archive at the University of Reading

Reading's research outputs online

1 Impact of plant choice on rainfall runoff delay and reduction by hedge species

2 T. Blanusa<sup>1,2,\*</sup>, J. Hadley<sup>2</sup>

3

4 <sup>1</sup> Science Department, Royal Horticultural Society, Garden Wisley, Woking GU23

5 6QB, Woking, UK

6 <sup>2</sup> School of Agriculture, Policy and Development, University of Reading, RG6 6AS,

7 UK

8 \* Corresponding author: [tijanablanusa@rhs.org.uk](mailto:tijanablanusa@rhs.org.uk)

9

10 Abstract

11

12 Soil sealing and a decrease in vegetation cover in urban areas increase the  
13 likelihood and frequency of localised flooding. Populating the remaining green areas  
14 with vegetation, which can efficiently capture excess rainfall, is therefore important.  
15 We argue that urban hedges can be a useful tool in mitigating rainfall, so the  
16 understanding of optimal plant choice, and underlying traits which enable most rain  
17 attenuation, is needed.

18 We tested the hypothesis that higher plant evapo-transpiration rates and larger  
19 canopy size can be linked with reduced rainfall runoff in urban hedge species. We  
20 first characterised seven hedge species grown in individual containers. These were  
21 both deciduous and evergreen species, with a range of inherent canopy sizes and  
22 water requirements. We assessed their plant water use, leaf stomatal conductance,  
23 canopy rainfall retention, and runoff delay and reduction capacity. The species  
24 showing highest and lowest potential for runoff reduction were then investigated for  
25 their outdoor performance, when planted in a hedge-like form. Our findings suggest  
26 that – after three days between rainfall events - species such as *Cotoneaster* and  
27 *Crataegus* with larger and wide canopies, and with high evapo-transpiration / water  
28 use rates, delayed the start of runoff (by as much to 10-15 minutes compared to bare  
29 substrate) as well reduced the volume of rainfall runoff. For example, <5% of the  
30 applied rainfall had runoff with *Cotoneaster* and *Crataegus*, compared with >40% in  
31 bare substrate. Substrate moisture content at the time of rainfall (which is linked to  
32 plants' ET rate) was the key explanatory variable.

33

34 Additional keywords: *Cotoneaster*, *Crataegus*, flood mitigation, hawthorn, *Thuja*

35

## 36 **Introduction**

37 Rapid urbanisation and an increase in sealed surfaces due to paving over (Smith  
38 2010; Verbeeck et al. 2011) can be linked to higher incidences of localised flooding  
39 in urban areas (Perry and Nawaz 2008; Warhurst et al. 2014). However,  
40 appropriately chosen and well managed vegetation in different forms of green  
41 infrastructure (GI) can play a role in reducing flood risks. This includes domestic  
42 gardens (Cameron et al. 2012; Kelly 2016; Warhurst et al. 2014), street trees (Xiao  
43 and McPherson 2002), vegetation strips such as grass verges (Charlesworth 2010),  
44 as well as urban hedgerows and garden hedges (O'Sullivan et al. 2017). All these  
45 green areas help rainfall management chiefly through maintaining soil, as the main  
46 natural water store in urban areas (Pit et al. 1999). Presence of vegetation also  
47 increases the soil's ability to receive subsequent rainfall through increasing soil's  
48 water-storage capacity by water loss via evapo-transpiration (Stovin et al. 2012).  
49 Additionally, plant roots can improve soil structure and increase porosity, increasing  
50 drainage and soil's water-holding capacity (Bartens et al. 2008; Mueller and  
51 Thompson 2009). There is also an element of rainfall interception and retention in  
52 the canopy, thus delaying runoff (Crockford and Richardson 2000).

53 In the UK, domestic gardens in urban areas take up a significant proportion of urban  
54 footprint (15-25%, Cameron et al. 2012; Gaston et al. 2005). Garden hedges are a  
55 ubiquitous feature of UK front gardens and can thus provide a number of frontline  
56 services including rainwater capture and localised flood protection. A recent survey  
57 by the Royal Horticultural Society (RHS) suggested that the vegetated area of front  
58 gardens across the UK has decreased by as much as 15% in the period 2005-2015.  
59 Additionally, one in four UK front gardens are paved-over and nearly one in three  
60 front gardens have no plants (Anon 2016). We argue therefore, that maintaining

61 unsealed surfaces in domestic gardens, including features such as garden hedges,  
62 can reduce the flooding risks for domestic households and streets/neighbourhoods.  
63 The question, however, is to what extent can we maximise canopy capture and  
64 runoff reduction by careful plant species choice, with traits to maximise this service?  
65 Previous work in our group (Blanusa et al. 2015; Blanusa et al. 2013; Cameron and  
66 Blanuša 2016; Vaz Monteiro et al. 2017; Kemp et al. 2017) provides evidence for the  
67 notion that differences in plant structure and the rate/mode of physiological function  
68 lead to differences in the provision of various ecosystem services by urban  
69 vegetation. E.g. plants with larger leaf areas, lighter leaf colour and greater rates of  
70 evapo-transpiration (ET) provide greater extent of building and ambient cooling by  
71 green roofs, by reducing soil heat flux and increasing latent heat fluxes (Vaz  
72 Monteiro et al. 2017). Larger leaf areas and greater ET rates of vegetation on green  
73 roofs have also been linked to reduced rainfall runoff rates (Kemp 2018).

74 Recent work on urban hedgerows (O'Sullivan et al. 2017) suggested that species  
75 with high water use are more efficient at reducing flooding risks. Ranking of species  
76 in that study is based on Roloff et al. (2009) work on drought tolerant trees (i.e.  
77 O'Sullivan et al. (2017) assume that less drought tolerant species have higher water  
78 use and thus offer greater flood protection). While this is a logical principle, no  
79 practical testing of hedge species had been carried out to explore this in practice. In  
80 the urban setting, other green infrastructure installations such as rain gardens or  
81 bioswales, and green roofs have been extensively studied for their capacity to  
82 reduce rainfall runoff (Berretta et al. 2014; Cameron and Hitchmough 2016;  
83 Scharenbroch et al. 2016), but the role of hedgerows in rainfall mitigation has been  
84 understudied.

85 A small body of existing work investigating rainfall management and runoff reduction  
86 by hedgerows was focused on rural / agricultural landscapes, rather than urban  
87 areas (e.g. Ghazavi et al. 2008; Herbst et al. 2006). Study by Herbst et al. (2006)  
88 quantified the rainfall interception loss of agricultural hedgerows per unit ground  
89 area, and determined the horizontal extension of the zone which is being influenced  
90 by the presence of a hedgerow. Two hedgerows in this study were composed  
91 predominantly of *Crataegus monogyna* (hawthorn), with some *Acer campestre* (field  
92 maple) sections, so the emphasis was on determining a general hedge effect rather  
93 than distinguishing the contribution of two species. Over the course of nearly a year  
94 these hedgerows intercepted >50% of the rainfall falling on the projected canopy  
95 area (Herbst et al. 2006). The width of the zone where hedges reduced runoff was  
96 equivalent to approximately two hedgerow heights and runoff reduction, during the  
97 period of full leaf cover, was 24% (Herbst et al. 2006). This is comparable to the  
98 highest observed values for a similar area of broadleaf tree stands and just slightly  
99 lower than coniferous woods (Herbst et al. 2006).

100 In addition to work on hedges' rainfall mitigation in agricultural context, a number of  
101 studies focus on individual tree specimens of species which could also be utilised as  
102 hedges (Keim et al. 2006; Nordén 1991; Asadian and Weiler 2009). Even so, very  
103 few potential hedge species have been studied in terms of the rainfall interception /  
104 retention e.g. *Thuja plicata* (Keim et al. 2006), *Fagus sylvatica* and *Carpinus betula*  
105 (Nordén 1991). These studies found *Thuja* had low capacity for water storage within  
106 the canopy compared to broad-leaved tree species (e.g. *Acer sp.*, *Rubus sp.* etc),  
107 but similar to other coniferous trees (e.g. *Tsuga heterophylla*, Keim et al., 2006). As  
108 a general guide, branches of all tree species tested in that study retained more water  
109 at higher, rather than lower rainfall intensities; leaf area was the best predictor of

110 canopy water storage, but more strongly for broadleaved than for needle-leaved  
111 species (Keim et al. 2006).

112 The aim of our study was therefore to test a range of urban hedge species (both  
113 deciduous and evergreen) differing in inherent vigour and canopy sizes, and with  
114 varying water use requirements and evapo-transpiration rates. We hypothesised that  
115 species exhibiting higher evapo-transpiration rates, which lead to a reduction in soil  
116 moisture content, can be linked with reduced rainfall runoff. We also hypothesized  
117 that species with larger canopy would exhibit greater runoff reduction. Our approach  
118 was two-pronged. We first characterised individual plant specimens of the selected  
119 species: their plant water use, leaf stomatal conductance, canopy rainfall retention,  
120 and runoff delay and reduction capacity. We have then selected the species showing  
121 highest and lowest runoff reduction and investigated their outdoor performance,  
122 when planted in a hedge-like form. Our findings suggest that the species with high  
123 water use rates, which reduced substrate moisture more before the rainfall was  
124 applied, better delayed the start of runoff as well reduced the volume of runoff.

125

## 126 **Materials and methods**

### 127 Rainfall application setup

128 To simulate natural rainfall in a controlled and repeatable manner, a sprinkler system  
129 based on the design described by Iserloh et al. (2012), produced 'in house' by an  
130 irrigation specialist at RHS Garden, Wisley, was used. The system consisted of a  
131 Lechler 460 608 nozzle attached to a 2 m length of hosing (Tricoflex, Hozelock Ltd.,  
132 Birmingham, UK) to a flow control, which was a series of pressure gauges and filters  
133 that ensured that the water flow and the characteristics of the droplets produced



134 were constant. The system was connected to the mains water supply by hosepipe,  
135 and rainfall could be turned on and off directly on the simulator (Figure 1). The  
136 optimum flow pressure to achieve consistent rainfall in terms of droplet size and  
137 distribution was found to be 0.15 bars (15 kPa), and so this pressure setting was  
138 used for all rainfall simulations. The nozzle, hosing and simulator were fastened to  
139 an L-shaped timber support structure 2.4 m high and 1 m across; this was then  
140 secured to a pre-existing metal frame in both glasshouse and field set-up, which run  
141 above all containers or troughs in the experiment.

142 The height of the nozzle was 0.7-0.9 m above the top of the experimental  
143 containers/troughs, depending on the height of the canopy in different species; this is  
144 in line with the heights of other rainfall simulators cited in the literature, typically for  
145 used in soil erosion and runoff studies, which vary between 0.7 and 3 m above the  
146 ground (e.g. Humphry et al. 2002; Fister et al. 2012). To further characterise the  
147 simulated rainfall, average raindrop size was measured using the flour pellet method  
148 described by Clarke and Walsh (2007). The diameters of all raindrops in three  
149 representative 4 x 4 cm areas were then measured using Image J software (National  
150 Institutes of Health, USA). Raindrop sizes ranged from 0.21 to 2.76 mm with the  
151 majority of droplets (70%) smaller than 1 mm diameter, similar to the simulated  
152 raindrops produced in other studies (e.g. Iserloh et al. 2012; Fister et al. 2012).

### 153 Experiments with individual hedge plants

154 Experiments were carried out in the period May-June 2016 in the ventilated  
155 glasshouses at the University of Reading (UK), where temperatures were maintained  
156 in the range 23-25 °C during daytime and 17-18 °C at night-time, with ambient light  
157 levels Four-year-old plants of seven hedge species, grown individually in 10 l

158 containers, with John Innes no 3 compost (7:3:2 sterilised loam:peat:coarse sand  
159 v/v, Westland, Dungannon, UK), were used. Species included five evergreen:  
160 *Photinia x fraseri* (cv 'Red Robin'), *Thuja plicata* (cv. 'Atrovirens'), *Taxus baccata*,  
161 *Ligustrum ovalifolium* (cvs. 'Aureum' and 'Argenteum') and *Cotoneaster franchetii*, as  
162 well as two deciduous species: *Crataegus monogyna* and *Fagus sylvatica*. Six  
163 replicates of each species were used, along with three containers with just bare  
164 substrate.

165 Two types of experiments were carried out. One was measuring contribution of  
166 canopy to runoff reduction (so carried out on plants immediately after the substrate  
167 was saturated to full container capacity ( $> 0.40 \text{ m}^3 \text{ m}^{-3}$ ). The other was measuring  
168 the importance of substrate moisture content and different ET rates to runoff  
169 reduction, by rainfall applications 3 days post saturation, with no additional watering  
170 in the 3 day period.

171 At the start of the experiment all containers were watered to full container capacity.  
172 Rainfall was applied either for 20 minutes (when measuring canopy interception, in  
173 containers where substrate was fully water-saturated) or 40 minutes (when  
174 containers were not watered for 3 consecutive days). Before simulated rainfall  
175 application, plant containers were placed within another 'collection' container which  
176 closely fitted but was 10 cm deeper, so that only the runoff from the substrate can be  
177 collected. To determine the runoff from each of the plant containers, water volume  
178 collected within the 'collection' container was measured after plants were left to drain  
179 for 1 h after the 'rainfall' stopped. For all rainfall applications, the rainfall simulator  
180 was fixed in a same position on a pre-existing metal frame within the glasshouse  
181 compartment. Position of the containers underneath the rainfall simulator was  
182 established by prior tests with 54 empty buckets (Kemp 2018, Kemp et al 2018) to

183 determine the uniformity of rainfall application and volume of applied rainfall. The  
184 positions underneath the rainfall simulator nozzle which provided an average volume  
185 of  $28 \pm 0.9 \text{ mm h}^{-1}$  were chosen. Additionally, we determined the volumes of water  
186 captured in containers of various diameters ( $d = 28 \text{ cm}$ ,  $41 \text{ cm}$  and  $69 \text{ cm}$ , all  
187 circular, plus a  $100 \times 100 \text{ cm}$  tray). The mean volumes of rainfall (from 2 tests)  
188 captured after a 40 m simulated rainfall event in these trays were 820 ml, 1100 ml,  
189 3145 and 8500 ml (on order of progressing size) That enabled us to calculate  
190 volumes of water received by canopies of various diameters and with different  
191 horizontal canopy projections. Once the experiment started, simulated rainfall for all  
192 replicate plants within one species would have been applied during the same day, by  
193 testing three and then two individual containers in the pre-determined fixed positions  
194 below the nozzle. As we had 7 species/cultivars to test in each experimental run, two  
195 days were required to test all species/plants. In testing the canopy retention,  
196 substrate was fully saturated just before the start of the experiment on each  
197 occasion, so the timing of rainfall application would have made no difference to the  
198 outcome. If testing the contribution of ET, the fact that experimentation was carried  
199 out over two days was mitigated by adding the water lost in the first 24 h cycle (as  
200 determined by weighing the plants) to the containers which would have been  
201 measured on the later day, so that altogether all plants experienced 72 h of ET loss  
202 at the moment of testing.

203 Before the start of the experiment, canopy width was determined by taking two  
204 perpendicular measurements. This was so that we can calculate plants' horizontal  
205 canopy projection which is capturing, and funnelling, rainfall and thus estimate the  
206 volumes of water which each canopy received. Wider canopies are exposed to - and  
207 have a potential to 'catch' - more water, so they could produce more runoff. We

208 therefore expressed our runoff data as a % of runoff water relative to the volume of  
209 rainfall received, in addition to absolute values of runoff volume. Additionally, plant  
210 height was measured, so that the canopy volume could be calculated from height  
211 and width measurements.

212 Measured parameters relating to canopy's capacity to capture rainfall included the  
213 weight of the containers with plants before and after rainfall application; this enabled  
214 us to quantify the weight of rainfall retained on the canopy in the situation when soil  
215 was fully saturated, as all the weight increase would be a result of what is held in the  
216 canopy (Eq 1).

$$217 \quad C_s = W_r - W_s \text{ (Eq 1)}$$

218 where:  $C_s$  - canopy rainfall storage capacity,  $W_r$  - weight of a plant and saturated  
219 container at the end of rainfall application,  $W_s$  - weight of a plant and saturated  
220 container just before rainfall application.

221 We also measured the substrate moisture content (SMC) using a SM300 sensor  
222 connected to a HH2 Moisture Meter (Delta-T Devices Ltd., Cambridge, UK) in two  
223 locations per container.

224 All species were then left for 72 h without watering and all containers were weighed  
225 daily using a precision balance (CBK 32, Adam Equipment, Milton Keynes,  
226 Buckinghamshire, UK), to estimate daily evapo-transpiration (ET) by plants and bare  
227 substrate. Substrate moisture content (SMC) was also recorded daily. After this 72 h  
228 period without watering, plants were subjected to second simulated rainfall and the  
229 volume of rainfall runoff was recorded. In doing that, we investigated the impact of  
230 plant ET and different rates of substrate drying in different species, on the volume of  
231 rainfall runoff. Both canopy sequestration and ET contribution experiments were

232 repeated twice over a two week period with different species tested in random order  
233 on the two occasions to minimise the impact of slight possible environmental  
234 differences in the glasshouse compartment on different days. Runoff data from both  
235 repeats matched closely, so only the data from the second repeat are shown in this  
236 paper.

237 Leaf stomatal conductance to water vapour was measured (using AP4 porometer,  
238 Delta-T Devices Ltd., Cambridge, UK) twice during the experiment: at the start of the  
239 experiment when plants were well-watered (i.e. substrate moisture content  $> 0.30 \text{ m}^3$   
240  $\text{m}^{-3}$  and also at the end of the experiment when the substrate was allowed to dry ( $<$   
241  $0.20 \text{ m}^3 \text{ m}^{-3}$ ). All treatments were measured on the same day in random order; three  
242 young fully expanded leaves per plant on five plants per species were used.

243 Additionally, at the end of the experiment, leaf area was measured destructively on  
244 three plants per species (apart from *Fagus* and *Crataegus* which were not  
245 measured) using a WinDIAS leaf area meter (Delta-T Devices, Cambridge, UK).

#### 246 Experiments with model hedges in troughs

247 Experiments were carried out in the period May-June 2017 on the outdoor field plots  
248 within the glasshouse complex at the University of Reading (UK). Five year old  
249 plants of *Crataegus monogyna* (common name: hawthorn), *Cotoneaster franchetii*  
250 and *Thuja plicata* (common name: yew) were transplanted from 10 L into 110 L  
251 troughs (1 m (l) x 0.4 m (w) x 0.45 m (d)) with Sylvamix substrate (6:2:2 sylvafibre:  
252 growbark pine: coir v/v; Melcourt, Tetbury, UK) with a slow-release fertiliser feed  
253 (Osmocote, Scotts, Marysville, OH, USA) in March 2017. There were three plants  
254 per container and three containers per species, along with three containers with just  
255 bare substrate ('control').

256 Before transplanting, each container was lined with a double layer of fine horticultural  
257 mesh (Veggiemesh Insect Netting, 1.35 mm mesh size) to aid retention of small  
258 substrate particles and prevent blockage of drainage holes. Mesh was then covered  
259 with 10 L of horticultural gravel (size 10 mm), followed with 80 L of substrate.

260 Plants were maintained outdoors and watered as required. Two weeks before the  
261 start of rainfall experiments, plants were cut into a hedge shape; *Thuja* and  
262 *Crataegus* were 1.1 m wide and *Cotoneaster* 1.2 m. Height and depth dimensions for  
263 each species are shown in Table 1. Height and depth measurements were made on  
264 three sections per trough, for each of the troughs at the start of the experiment.

265 Indicative leaf area for each species was determined destructively at the end of the  
266 experiment by cutting out two 15 cm x 15 cm x 15 cm sections in each replicate of  
267 the model hedges and measuring with leaf area meter (Delta-T Devices, Cambridge,  
268 Cambridgeshire, UK).

269 [Insert Table 1]

270 At the beginning of the experiment, troughs with hedge plants and bare soil were put  
271 into fixed positions in a field plot. The twelve experimental troughs were arranged in  
272 two parallel rows of six; arrangement of troughs within a row was random. Each  
273 trough was placed onto a plastic tray (1.1 m (l) x 0.45 m (w) x 0.05 m (d)) and both  
274 were then elevated onto a pedestal at 4° angle, constructed from bricks and wood  
275 planks; this enabled the water to drain freely through the holes drilled on one end of  
276 the tray. During the experiment, to collect the rainfall runoff, plastic containers were  
277 fitted under the tray holes. Experimental setup is shown in Figure 1.

278 [Insert Figure 1]

279 The time taken for runoff to be generated from trays with bare substrate was pre-  
280 tested with the chosen rainfall simulator settings, and found to vary between 5 and  
281 15 minutes, depending on initial substrate moisture content. As the plants would be  
282 increasing rainfall retention, to ensure that measurable runoff was always generated  
283 from all planted treatments and all substrate moisture conditions, it was therefore  
284 decided to simulate rainfall for 20 minutes (for troughs saturated to full water-holding  
285 capacity, where the role of canopy retention in runoff reduction was measured) or 60  
286 minutes (for troughs after 3 days without irrigation, where the role of ET in runoff  
287 reduction was measured) for each container/trough (Table 2).

288 To set up the rainfall applicator, on the ground, at the back of the trough, a fixed  
289 position for the timber support and rainfall applicator was marked at the same  
290 distance from each trough, so all rainfall applications were administered from the  
291 same location for each trough.

292 Since rainfall could only be applied to one trough at a time, this meant that only 8  
293 troughs could be tested in a working day (when the 60 min application time and  
294 subsequent draining times were factored in). Each experimental run was therefore  
295 conducted over two consecutive days, testing two replicates from each treatment on  
296 day 1 and one replicate on day 2. Experimental runs were carried on relatively still  
297 days, with wind speed  $< 5 \text{ m s}^{-1}$ .

298 Two types of experiments were carried out (Table 2). One was measuring  
299 contribution of canopy (so carried out on hedges where the substrate is saturated to  
300 full container capacity ( $> 40 \text{ m}^3 \text{ m}^{-3}$ )). The other was measuring the importance of  
301 substrate moisture content and different ET rates for runoff reduction, by rainfall  
302 applications after 3 days post-saturation. Due to the treatments' different ET rates,

303 this would have led to different starting SMCs for this experiment. Details of  
304 measurements are shown in Table 2.

305 **[Insert Table 2]**

306 At the start of the first experiment all containers were watered to full capacity.  
307 Experiments were repeated three times in a four week period and all data was  
308 analysed together as described in the Statistics section.

309 Before the start of the rainfall runoff experiments, a baseline measurement of leaf  
310 stomatal conductance to water vapour and net CO<sub>2</sub> assimilation of each plant  
311 treatment was made to establish plants' ET capacity, when substrate moisture  
312 content is at the field capacity. Three young fully expanded leaves per plant, on  
313 every plant, in two troughs per species (i.e. 9 measurements per trough, 18 per  
314 species) were measured using LCpro infra-red gas analyser (ADC Bioscientific,  
315 Hoddesdon, UK).

316 Before each simulated rainfall run, substrate moisture content in each trough was  
317 measured using a SM300 sensor connected to a HH2 Moisture Meter (Delta-T  
318 Devices Ltd., Cambridge, UK) in four locations per trough.

### 319 Statistical analysis

320 For experiments with individual containers, an analysis of variance (ANOVA) was  
321 performed using GENSTAT (18th Edition, VSN International, Hemel Hempstead,  
322 Hertfordshire, UK). There, we compared means for each measured parameter  
323 (runoff volumes, canopy retention, leaf stomatal conductance, water loss by plants  
324 etc.) between different species. Variance levels were checked for homogeneity and  
325 values were presented as means with associated least significant differences, which



326 were used to assess variations at a 5% significance level. Additionally, linear  
327 regression analysis was performed to establish a relationship between parameters  
328 such as ET and  $g_s$ , and runoff volumes.

329 For the experiments with hedges in troughs, to analyse runoff volumes from three  
330 consecutive sets of experiments, a repeated measurements analysis was employed.  
331 Linear mixed models were used to model the relationship of responses with the  
332 explanatory factors and covariates. The response 'runoff volume' was modelled on a  
333 logarithmic scale, hence its effect measures are expressed in Results tables as  
334 ratios of predicted means. 'Species' and 'minutes after rainfall application ceased'  
335 were fitted as fixed effects; 'date' and 'trough' were fitted as random effects to make  
336 results from this experiment more generalizable to users. To account for the  
337 correlated measurements taken on the same trough over time, an unstructured  
338 marginal covariance structure was used for the term 'minutes after rainfall application  
339 ceased'. All overall F-test were adjusted using a Kenward-Roger method in PROC  
340 MIXED of SAS version 9.4. Finally, post-modelling pairwise comparisons between  
341 species were adjusted for multiplicity using a Holm method. For the analysis of  
342 substrate moisture content within troughs, net leaf CO<sub>2</sub> assimilation and leaf  
343 stomatal conductance on individual dates, a one-way ANOVA was performed as  
344 described for individual containers.

345

## 346 **Results**

### 347 Experiments with individual hedge plants

348 In our experiment, *Photinia* 'Red Robin' had the largest canopy leaf area (1.64 m<sup>2</sup>),  
349 with all other species being statistically similar and averaging around 0.65 m<sup>2</sup> (data

350 not shown). The branch orientation and crown horizontal canopy ground projection  
351 differed between the species, with *Cotoneaster* and *Photinia* having largest and  
352 *Thuja* having lowest canopy ground projection (Table 3). Canopy volume was  
353 greatest for *Cotoneaster* and *Photinia* and lowest for *Thuja* (Table 3). Plant heights  
354 however, were mostly similar between species (averaging 113 cm) with just *Photinia*  
355 being significantly taller, at 143 cm (data not shown).

356 Canopy retention of the rainfall was greatest in the two *Ligustrum* cultivars  
357 (averaging close to 400 ml per plant), and lowest in *Thuja* (below 250 ml per plant),  
358 with other species being similar at around 310 ml per plant (Table 3). Linear  
359 regression analysis revealed no statistically significant relationship between canopy  
360 volume and canopy retention ( $p = 0.19$ ).

361 [Insert Table 3]

362 Leaf stomatal conductance (measured when plants were well watered, on Day 1 of  
363 the experiment) was highest in *Cotoneaster* and *Crataegus* (around  $200 \mu\text{mol m}^{-1} \text{s}^{-1}$   
364  $^{-1}$ ) and lowest in *Thuja* and *Taxus* (below  $100 \mu\text{mol m}^{-1} \text{s}^{-1}$ ) (Table 3). *Cotoneaster*,  
365 *Crataegus* and *Photinia* lost most water per plant (over 2000 ml in a in 72 h period)  
366 with *Thuja* losing least of all plant treatments (<1500 ml). All plant treatments lost  
367 significantly more water than just bare soil (just over 600 ml in a 3-day period) (Table  
368 3).

369 Substrate moisture content after 3 days with no irrigation was lowest in *Cotoneaster*  
370 ( $0.20 \text{ m}^3 \text{ m}^{-3}$ ) and highest in bare substrate ( $0.45 \text{ m}^3 \text{ m}^{-3}$ ); all other plant treatments  
371 had SMC between  $0.28$  and  $0.30 \text{ m}^3 \text{ m}^{-3}$  (data not shown). Canopies of different  
372 species have different spreads, and thus different ground projections (Table 3).  
373 Water volumes received by different canopies are thus also different (Table 4).

374 [Insert Table 4]

375 Runoff from the containers, where rainfall was applied after 3 days with no watering,  
376 was negligible from *Crataegus* both in absolute terms (Table 4), and when  
377 expressed relative to the volume of water received (Figure 2). *Cotoneaster* too had  
378 lower volume of runoff (when rained on after 3 days with no watering) compared to  
379 all other species (apart from *Crataegus*, relatively expressed) (Figure 2). In absolute  
380 terms, but also in relation to the volume of rainfall received, *Thuja* had the highest  
381 runoff off all the plant species, although it was still lower than for the bare substrate  
382 (Table 4, Figure 2).

383 [Insert Figure 2]

384 Linear regression analysis revealed no statistically significant relationship between  
385 ET or  $g_s$ , and runoff volumes (data not shown). There was a statistically significant ( $p$   
386 = 0.05) positive linear relationship between SMC and runoff volume (when  
387 expressed as a % volume received) ( $R^2 = 0.14$ ).

388

389 Experiments with model hedges in troughs

390 Substrate moisture content was similar for all the treatments at the start of the  
391 experiment, then lower in all plant treatments after 3 and 5 days of drying compared  
392 bare soil (Table 5). Additionally, net CO<sub>2</sub> assimilation and leaf stomatal conductance  
393 were statistically significantly higher, when measured on Day 1 of the experiment in  
394 well-watered *Cotoneaster* than in *Crataegus* and *Thuja* (Table 5).

395 [Insert Table 5]

396 When substrate was fully saturated (i.e. only the canopy provided the barrier to  
397 rainfall), runoff was recorded first from a bare substrate treatment, then *Thuja*  
398 followed by *Cotoneaster* and *Crataegus* (Table 6A); statistical analysis showed  
399 significant treatment differences (P = 0.032, data not shown). *Cotoneaster* and  
400 *Crataegus* delayed runoff longer than bare substrate (Holm p-values 0.055 and  
401 0.051, respectively). Statistical analysis showed no significant influence of either  
402 canopy volume or canopy density on the delay of runoff (p = 0.3669 and 0.6167,  
403 respectively) (data not shown).

404 In terms of volumes of runoff after the rain stopped falling on previously saturated  
405 substrate there were significant treatment differences. The volume of runoff  
406 generated at the end of rainfall was greatest in bare soil and *Thuja*, least in  
407 *Cotoneaster* and *Crataegus* (Table 6B). *Cotoneaster* and *Crataegus* produced  
408 statistically significantly less runoff than bare soil (e.g. at the end of the rainfall: Holm  
409 p-values 0.010 and 0.013 respectively).

410 [Insert Table 6]

411 After three days with no irrigation, substrate moisture content was on average 0.27,  
412 0.18, 0.17 and 0.18 m<sup>3</sup> m<sup>-3</sup> for bare soil, *Thuja*, *Crataegus* and *Cotoneaster*

413 respectively (Table 5). Statistically, at that time point all plant species had similar  
414 substrate moisture, and all statistically lower than bare soil (Table 5).

415 When rainfall was applied to treatments after 3 days of no irrigation there were  
416 significant treatment differences in terms of the extent of runoff delay. There was a  
417 significant species effect ( $p = 0.0110$ ) in the delay of runoff, with both *Cotoneaster*  
418 and *Crataegus* delaying runoff more than bare substrate and *Thuja* (Table 7A). In  
419 terms of volumes of runoff there were again significant species differences ( $p =$   
420  $0.0258$ ). Particularly, after 60 min draining there was significantly less runoff from  
421 *Crataegus* and *Cotoneaster* compared to bare substrate ( $p = 0.0083$ ) (Table 7).

422 [Insert Table 7]

423 Statistical analysis showed the significant influence of substrate moisture content on  
424 both delay of runoff and the volumes of runoff ( $p = 0.0397$  and  $0.0551$ , respectively),  
425 but there was no impact of leaf stomatal conductance ( $p = 0.5414$  and  $0.4470$ ,  
426 respectively).

427

## 428 Discussion

429 Loss of vegetation in urban areas, and in domestic gardens (in the UK) in particular  
430 can be linked to higher incidences of localised flooding in urban areas (Perry and  
431 Nawaz 2008; Warhurst et al. 2014). In a context of most domestic households in the  
432 UK having their own domestic garden (Cameron et al., 2012), urban hedges as a  
433 ubiquitous garden feature could be seen as a frontline protection for households  
434 from localised flooding. This is due to the delay of rainfall runoff when rainfall is  
435 captured on the canopies (i.e. canopy interception) and absorbed into the soil. With  
436 front gardens and associated hedges increasingly being lost to paving, making sure

437 that the hedges we do plant and retain are providing maximal rainfall attenuation is  
438 important. We argue that through careful choice of hedge species, rainfall mitigation  
439 by urban hedges can be maximised.

440 Previous research found that depending on the intensity of the rainfall, canopy  
441 capture (e.g. in juniper trees) can represent 20-60% of bulk precipitation, with more  
442 canopy capture in less intense events (Carlyle-Moses, 2004, Owens et al., 2006).  
443 Additionally, in a young sitka spruce plantation, canopies captured 30% of rainfall  
444 annually (Ford and Deans 2018). Rainfall captured and temporarily retained in the  
445 canopy is especially important in a scenario of rainfall events happening in close  
446 sequence, when there is insufficient time for ET (particularly plants' transpirational  
447 component which removes water from the soil) to make a significant contribution to  
448 runoff reduction. Characteristics such as area covered by vegetation, branch angle,  
449 the uniformity in crown height, nature of the bark, leaf shape and inclination, and leaf  
450 area index will all influence rainfall interception by the canopies (Crockford and  
451 Richardson 2000). Branch diameter was also found to be positively correlated with  
452 canopy rainfall retention in several forest coniferous species (Liu 1998). Additionally,  
453 factors such as intensity of rainfall and other meteorological conditions (temperature,  
454 humidity, wind speed etc.) will have a role (Crockford and Richardson 2000, Toba  
455 and Ohta, 2005). While the conclusions about the contribution of various factors to  
456 rainfall capture and runoff reduction are generated largely from the forest and  
457 individual trees literature, they none the less present a starting point in interpreting a  
458 role that different hedges' forms and function might have in these processes. Due to  
459 the smaller area they cover, impact of hedges, of course will be more localised e.g.  
460 affecting an individual garden rather than a street-level catchment.

461 In our experiment, although species with greater leaf area (e.g. *Ligustrum*) generally  
462 retained more rainfall in the canopy, this was not always the case (e.g. *Photinia*). In  
463 our experiment just one rainfall intensity was tested; a response of different canopy  
464 structures to a change on rainfall intensity might vary (Carlyle-Moses and Gash,  
465 2011). Based on our measurements, canopy leaf area, or even canopy volume, were  
466 clearly not the only explanatory variables of canopy retention, with species having  
467 similar leaf areas but different canopy retentions (e.g. *Ligustrum* vs *Taxus* or *Thuja*).  
468 While we could not numerically capture all the possible parameters potentially  
469 influencing canopy retention, the presence of clear species differences and  
470 anecdotal observations within our experiment would suggest that factors such as  
471 dense or more horizontal branch architecture, concave leaf shape and presence of  
472 structures like leaf hairs played a role in improving rainfall canopy capture.

473 While acknowledging the importance of canopy structural characteristics in rainfall  
474 retention, our primary interest was in establishing the contribution of plant functional  
475 characteristics such as ET and leaf stomatal conductance to runoff reduction. This  
476 was because of their impact on soil/substrate content which had been shown, in a  
477 green roof context at least, as an important predictor of rainfall runoff reduction  
478 (Kemp et al., 2018; Stovin et al. 2012; Poë et al. 2015).

479 Larger canopies receive more water into the canopy and filter it towards the ground  
480 (Ford and Deans, 1978). In our experiment, *Cotoneaster* covered the largest area  
481 over the ground, hence was exposed to most rainfall, yet had one of the lowest  
482 runoff rates. *Thuja*, conversely, has the smallest ground projection, but together with  
483 *Photinia* has highest runoff values amongst the studied species. Our observations in  
484 the outdoor experiment suggest that it was the branch architecture of *Thuja* (where  
485 branches are generally at 30-45° away from the trunk) which encouraged more water

486 to be funnelled towards the trunk and ultimately soil (causing more runoff), compared  
487 with species where branches and leaves are positioned closer to a 90°. This  
488 however could be seen as a positive on more free-draining soils (as it would channel  
489 more rainfall towards the ground). Conversely, *Cotoneaster* and *Crataegus* would  
490 offer best protection in soils which are less free-draining.

491 In both sets of experiments antecedent substrate moisture content was positively  
492 correlated with volumes of runoff. Our earlier preliminary experiment with the same  
493 species showed that *Cotoneaster* and *Crataegus* lost most water per m<sup>2</sup> of leaf area  
494 in any 24 h period (Blanusa et al. 2017) and they were the ones which then  
495 produced lowest runoff rates in subsequent experiments. In our experiment with  
496 hedges in troughs outdoors, runoff was lower in all plant treatments compared to  
497 bare substrate. This would thus suggest that lowering SMC and higher ET had some  
498 advantage in the first 2-3 days after the rainfall in an outdoor summertime scenario.

499 Individually, other functional parameters such as leaf stomatal conductance and ET  
500 were not statistically significantly linked to a delay or reduction of runoff. It is  
501 therefore likely that while low antecedent substrate moisture plays an important in  
502 delaying and reducing the runoff in hedge species, an additional complex  
503 combination of variables such as canopy shape and leaf properties (e.g. leaf  
504 hydrophobicity, Holder 2013) as well as root density and structure also play part.

## 505 **Conclusions**

506 Urban hedges are an important green infrastructure component in urban areas and  
507 particularly in people's domestic (front) gardens in the UK where they are a popular  
508 and, arguably, widely spread feature. They have a capacity to delay and reduce  
509 rainfall runoff and thus offer protection from localised flooding, within an urban



510 environment where loss of vegetated surfaces has been lined with increased  
511 incidents of flooding. Our experiments showed a significant impact of species choice  
512 on a hedge's capacity to retain water on the canopy, as well as to delay and reduce  
513 runoff. Of the studied species, *Ligustrum* and *Cotoneaster* retained largest rainfall  
514 volumes within their canopies. While we could not numerically capture all the  
515 possible parameters potentially influencing canopy retention, the presence of clear  
516 species differences and observations within our experiment suggest that factors  
517 such as dense or more horizontal branch architecture, concave leaf shape and  
518 presence of structures like leaf hairs played a role in improving rainfall canopy  
519 capture.

520 Hedge species such as *Cotoneaster* and *Cataegus*, delayed the start of runoff (by as  
521 much to 10-15 minutes compared to bare substrate) as well reduced the volume of  
522 rainfall runoff. For example, <5% of the applied rainfall had runoff with *Cotoneaster*  
523 and *Crataegus*, compared with >40% in bare substrate. Substrate moisture content  
524 at the time of rainfall (which is linked to plants' ET rate) seems to be the key  
525 explanatory variable.

526

## 527 **Acknowledgements**

528 The authors are grateful to Paul Mealey for designing and producing rainfall  
529 simulators, Kevin Hobbs at Hillier Nurseries for the supply of plants, Matthew  
530 Richardson, Will Johnson, Val Jasper, Julia Janes, Michael Dawes, Curtis Gubb for  
531 expert technical help, Alessandro Leidi for statistical support and Dr Sarah Kemp, Dr  
532 Andrew Daymond, Dr Paul Alexander and Leigh Hunt for constructive discussions.

533

534 **References**

- 535 Anon (2016) How green are British front gardens? Ipsos MORI. [https://www.ipsos-](https://www.ipsos-mori.com/researchpublications/researcharchive/3738/How-green-are-British-front-gardens.aspx)
- 536 [mori.com/researchpublications/researcharchive/3738/How-green-are-British-front-](https://www.ipsos-mori.com/researchpublications/researcharchive/3738/How-green-are-British-front-gardens.aspx)
- 537 [gardens.aspx](https://www.ipsos-mori.com/researchpublications/researcharchive/3738/How-green-are-British-front-gardens.aspx). Accessed 29 November 2016
- 538 Asadian Y, Weiler M (2009) A new approach in measuring rainfall interception by urban
- 539 trees in coastal British Columbia. *Water quality research journal of Canada* 44 (1):16
- 540 Bartens J, Day SD, Harris JR, Dove JE, Wynn TM (2008) Can urban tree roots improve
- 541 infiltration through compacted subsoils for stormwater management? *Journal of*
- 542 *Environmental Quality* 37 (6):2048-2057
- 543 Berretta C, Poë S, Stovin V (2014) Moisture content behaviour in extensive green roofs
- 544 during dry periods: The influence of vegetation and substrate characteristics. *Journal*
- 545 *of Hydrology* 511:374-386
- 546 Blanusa T, Fantozzi F, Monaci F, Bargagli R (2015) Leaf trapping and retention of particles
- 547 by holm oak and other common tree species in Mediterranean urban environments.
- 548 *Urban Forestry & Urban Greening* 14 (4):1095-1101.
- 549 doi:<http://dx.doi.org/10.1016/j.ufug.2015.10.004>
- 550 Blanusa T, Hadley J, Hunt L, Alexander P, Hobbs K (2017) Provision of ecosystem services
- 551 by hedges in urban domestic gardens: focus on rainfall mitigation. *Acta Horticulturae*
- 552 1189:519-523
- 553 Blanusa T, Vaz Monteiro MM, Fantozzi F, Vysini E, Li Y, Cameron RWF (2013) Alternatives
- 554 to Sedum on green roofs: Can broad leaf perennial plants offer better 'cooling
- 555 service'? *Building and Environment* 59: 99-106. doi:10.1016/j.buildenv.2012.08.011
- 556 Cameron R, Hitchmough J (2016) New green space interventions-green walls, green roofs
- 557 and rain gardens. In: *Environmental horticulture: science and management of green*
- 558 *landscapes*, CAB International, Boston, pp 260-283
- 559 Cameron RWF, Blanuša T (2016) Green infrastructure and ecosystem services – is the devil
- 560 in the detail? *Annals of Botany* 118 (3):377-391. doi:10.1093/aob/mcw129

561 Cameron RWF, Blanusa T, Taylor JE, Salisbury A, Halstead AJ, Henricot B, Thompson K  
562 (2012) The Domestic Garden - Its Contribution to Urban Green Infrastructure. *Urban*  
563 *Forestry & Urban Greening* 11 (2):129-137

564 Carlyle-Moses, D (2004) Throughfall, stemflow, and canopy interception loss fluxes in a  
565 semi-arid Sierra Madre Oriental matorral community. *Journal of Arid Environments*  
566 58 (2): 181-202.

567 Carlyle-Moses, DE, and Gash, JH (2011) Rainfall interception loss by forest canopies. In:  
568 Levia DF, Carlyle-Moses D, Tanaka T (eds) *Forest hydrology and biogeochemistry*,  
569 Springer, New York, pp 407-423

570 Charlesworth SM (2010) A review of the adaptation and mitigation of global climate change  
571 using sustainable drainage in cities. *Journal of Water and Climate Change* 1 (3):165-  
572 180. doi:10.2166/wcc.2010.035

573 Clarke MA, Walsh RPD (2007) A portable rainfall simulator for field assessment of splash  
574 and slopewash in remote locations. *Earth Surface Processes and Landforms* 32  
575 (13):2052-2069. doi:10.1002/esp.1526

576 Crockford R, Richardson D (2000) Partitioning of rainfall into throughfall, stemflow and  
577 interception: effect of forest type, ground cover and climate. *Hydrological Processes*  
578 14 (16-17):2903-2920

579 Fister W, Iserloh T, Ries JB, Schmidt RG (2012) A portable wind and rainfall simulator for in  
580 situ soil erosion measurements. *Catena* 91:72-84. doi:10.1016/j.catena.2011.03.002

581 Ford ED, Deans JD (1978) The Effects of Canopy Structure on Stemflow, Throughfall and  
582 Interception Loss in a Young Sitka Spruce Plantation. *Journal of Applied Ecology* 15  
583 (3):905-917. doi:10.2307/2402786

584 Gaston KJ, Warren PH, Thompson K, Smith RM (2005) Urban Domestic Gardens (IV): The  
585 Extent of the Resource and its Associated Features. *Biodiversity and Conservation*  
586 14 (14):3327-3349

587 Ghazavi G, Thomas Z, Hamon Y, Marie J-C, Corson M, Merot P (2008) Hedgerow impacts  
588 on soil-water transfer due to rainfall interception and root-water uptake. *Hydrological*  
589 *Processes* 22 (24):4723-4735

590 Herbst M, Roberts JM, Rosier PTW, Gowing DJ (2006) Measuring and modelling the rainfall  
591 interception loss by hedgerows in southern England. *Agricultural and Forest*  
592 *Meteorology* 141 (2–4):244-256.  
593 doi:<http://dx.doi.org/10.1016/j.agrformet.2006.10.012>

594 Holder CD (2013) Effects of leaf hydrophobicity and water droplet retention on canopy  
595 storage capacity. *Ecohydrology* 6 (3):483-490

596 Humphry JB, Daniel TC, Edwards DR, Sharpley AN (2002) A portable rainfall simulator for  
597 plot-scale runoff studies. *Applied Engineering in Agriculture* 18 (2):199-204

598 Iserloh T, Fister W, Seeger M, Willger H, Ries JB (2012) A small portable rainfall simulator  
599 for reproducible experiments on soil erosion. *Soil and Tillage Research* 124:131-137.  
600 doi:<http://dx.doi.org/10.1016/j.still.2012.05.016>

601 Keim R, Skaugset A, Weiler M (2006) Storage of water on vegetation under simulated  
602 rainfall of varying intensity. *Advances in Water Resources* 29 (7):974-986

603 Kelly D (2018) Impact of paved front gardens on current and future urban flooding. *Journal of*  
604 *Flood Risk Management* 11 (S1):434–443

605 Kemp S (2018) Impact of plant choice and water management on the provision of ecosystem  
606 services by green roofs. Dissertation, University of Reading, UK.

607 Kemp S, Blanusa T, Hadley P (2017) Greywater impact on green roofs' provision of  
608 ecosystem services. *Acta Horticulturae* 1189:513-518

609 Kemp, S, Hadley, P, and Blanuša, T (2019). The influence of plant type on green roof rainfall  
610 retention. *Urban Ecosystems*, 22 (2): 355–366.

611 Liu S (1998) Estimation of rainfall storage capacity in the canopies of cypress wetlands and  
612 slash pine uplands in North-Central Florida. *Journal of Hydrology* 207 (1-2):32-41

613 Mueller GD, Thompson AM (2009) The ability of urban residential lawns to disconnect  
614 impervious area from municipal sewer systems. *Journal of the American Water*  
615 *Resources Association* 45 (5):1116-1126

616 Nordén U (1991) Acid deposition and throughfall fluxes of elements as related to tree  
617 species in deciduous forests of South Sweden. *Water, Air, & Soil Pollution* 60  
618 (3):209-230

619 O'Sullivan OS, Holt AR, Warren PH, Evans KL (2017) Optimising UK urban road verge  
620 contributions to biodiversity and ecosystem services with cost-effective management.  
621 *Journal of Environmental Management* 191:162-171

622 Owens MK, Lyons RK, Alejandro CL (2006) Rainfall partitioning within semiarid juniper  
623 communities: effects of event size and canopy cover. *Hydrological Processes: An*  
624 *International Journal* 20 (15):3179-3189

625 Perry T, Nawaz R (2008) An investigation into the extent and impacts of hard surfacing of  
626 domestic gardens in an area of Leeds, United Kingdom. *Landscape and Urban*  
627 *Planning* 86 (1):1-13

628 Pit R, Lantrip J, Harrison R, Henry CL, Xue D (1999) Infiltration through disturbed urban soils  
629 and compost-amended soil effects on runoff quality and quantity. *National Risk*  
630 *Management Research Laboratory*. Washington, DC

631 Poë S, Stovin V, Berretta C (2015) Parameters influencing the regeneration of a green roof's  
632 retention capacity via evapotranspiration. *Journal of Hydrology* 523:356-367

633 Roloff A, Korn S, Gillner S (2009) The Climate-Species-Matrix to select tree species for  
634 urban habitats considering climate change. *Urban Forestry & Urban Greening* 8  
635 (4):295-308. doi:<https://doi.org/10.1016/j.ufug.2009.08.002>

636 Scharenbroch BC, Morgenroth J, Maule B (2016) Tree species suitability to bioswales and  
637 impact on the urban water budget. *Journal of environmental quality* 45 (1):199-206

638 Smith C (2010) *London: Garden city?* London Wildlife Trust, Greenspace Information for  
639 Greater London, Greater London Authority, London

640 Stovin V, Vesuviano G, Kasmin H (2012) The hydrological performance of a green roof test  
641 bed under UK climatic conditions. *Journal of Hydrology* 414:148-161

642 Toba, T, and Ohta, T (2005) An observational study of the factors that influence interception  
643 loss in boreal and temperate forests. *Journal of Hydrology* 313: 208-220.

644 Vaz Monteiro M, Blanuša T, Verhoef A, Richardson M, Hadley P, Cameron RWF (2017)  
645 Functional green roofs: Importance of plant choice in maximising summertime  
646 environmental cooling and substrate insulation potential. *Energy and Buildings*  
647 141:56-68. doi:<http://dx.doi.org/10.1016/j.enbuild.2017.02.011>

648 Verbeeck K, Van Orshoven J, Hermy M (2011) Measuring extent, location and change of  
649 imperviousness in urban domestic gardens in collective housing projects. *Landscape  
650 and Urban Planning* 100 (1):57-66.

651 Warhurst JR, Parks KE, McCulloch L, Hudson MD (2014) Front gardens to car parks:  
652 Changes in garden permeability and effects on flood regulation. *Science of the Total  
653 Environment* 485:329-339

654 Xiao Q, McPherson EG (2002) Rainfall interception by Santa Monica's municipal urban  
655 forest. *Urban Ecosystems* 6 (4):291-302

656

657 **List of Table captions**

658

659 Table 1. Mean hedge height and depth (in cm), as well as a mean indicative leaf  
660 area (in cm<sup>2</sup>) collected from a 15 cm x 15 cm x 15 cm a section within hedge canopy.  
661 Data are mean of two (leaf area) or three (height and depth) sections of hedge on  
662 each trough with associated least significant difference (LSD) between means ( $P <$   
663  $0.05$ ). Different letters next to the means in a column denote statistically significant  
664 difference between those means.

665

666 Table 2. Details of experimental conditions and measurements made in the outdoor  
667 experiment with model hedges in troughs.

668

669 Table 3. Average canopy volume, rainfall canopy retention, leaf stomatal  
670 conductance and ET, with the associated least significant differences between the  
671 means. Different letters next to the means in a column denote statistically significant  
672 difference between those means ( $P = 0.05$ ). Degrees of freedom (d.f.) are also  
673 shown.

674

675 Table 4. Mean rainfall volume received within a 40 minute event and volume of  
676 runoff. Least significant difference (LSD) and degrees of freedom (d.f.) are also  
677 shown. Different letters next to the means in a column denote statistically significant  
678 difference between those means ( $P = 0.05$ ).

679

680 Table 5. Mean substrate moisture content on days 1, 3, and 5 of the first  
681 experimental round (22-25 May 2017) along with net CO<sub>2</sub> assimilation and stomatal  
682 conductance values on day 1 when all plants were well watered. Least significant  
683 difference (LSD) and degrees of freedom (d.f.) are also shown. Different letters next  
684 to the means in a column denote statistically significant difference between those  
685 means ( $P = 0.05$ ); NS = non-significant.

686

687 Table 6. Predicted mean time to runoff (A) and runoff volumes (B) when rainfall was  
688 applied for 20 min onto troughs where substrate was fully saturated. Data are  
689 predicted means of three repeated experiments for all treatments and three troughs  
690 per treatment. Discussion of statistical significance in the body of the text is based on  
691 Holm p-values.

692

693 Table 7. Mean time to runoff (A) and runoff volumes (B) when rainfall was applied for  
694 60 min onto troughs where substrate was not watered for 3 days. Data are predicted  
695 means of three repeated experiments for all treatments and three troughs per  
696 treatment. Discussion of statistical significance in the body of the text is based on  
697 Holm p-values.

698

699



700 **List of Figure captions**

701

702 Figure 1. Setup for the outdoor experiment with model hedges in troughs.

703

704 Figure 2. Percent runoff in relation to the rainfall volume received per canopy, after a  
705 40 min simulated rainfall event with the intensity of 28 mm. Rainfall was applied 72 h  
706 after the plants were watered. Values are means of six replicates per plant species  
707 and three replicates for bare soil. Error bar represents least significant difference  
708 between the means (LSD,  $P = 0.05$ ).

709

710

711 Table 1. Mean hedge height and depth (in cm), as well as a mean indicative leaf  
 712 area (in cm<sup>2</sup>) collected from a 15 cm x 15 cm x 15 cm a section within hedge canopy.  
 713 Data are mean of two (leaf area) or three (height and depth) sections of hedge on  
 714 each trough with associated least significant difference (LSD) between means (P <  
 715 0.05). Different letters next to the means in a column denote statistically significant  
 716 difference between those means.

Species	Height (cm)	Depth (cm)	Leaf area (cm <sup>2</sup> ) within a 15 x 15 x 15 cm section of the canopy
<i>Cotoneaster</i>	73.3 a	120.4 a	801
<i>Crataegus</i>	51.8 b	114.0 a	1165
<i>Thuja</i>	151.1 c	61.2 b	1282
LSD	6.77 ***	15.65 ***	496.8 (ns)

717

718

719 Table 2. Details of experimental conditions and measurements made in the outdoor  
 720 experiment with model hedges in troughs.

Type of experiment	Watering and substrate moisture	Rainfall duration	Observations and measurements made				
			Time to runoff (min)	Volume of runoff at the end of the rainfall (ml)	Volume of runoff 20 min after rainfall end (ml)	Volume of runoff 60 min after rainfall end (ml)	Volume of runoff after 3 h (ml) = 'total'
Canopy interception	Watered to full container capacity before experiment start	20 min	X	X	X	X	X
Canopy and substrate interception	Not watered for 72 h prior to the start of experiment	60 min	X	X	X	X	X

721

722

723 Table 3. Average canopy volume, rainfall canopy retention, leaf stomatal  
 724 conductance and ET, with the associated least significant differences between the  
 725 means. Different letters next to the means in a column denote statistically significant  
 726 difference between those means ( $P = 0.05$ ). Degrees of freedom (d.f.) are also  
 727 shown.

Treatment	Canopy volume (m <sup>3</sup> )	Canopy retention (ml)	Canopy ground projection (m <sup>2</sup> )	Leaf stomatal conductance ( $\mu\text{mol m}^{-1} \text{s}^{-1}$ )	ET per plant in a 72 h period (ml)
Soil	-	-		-	627 e
<i>Thuja</i>	0.352 c	245 d	0.30 e	90.8 de	1465 d
<i>Taxus</i>	0.393 bc	280 cd	0.35 de	67.2 e	1917 bc
<i>Crataegus</i>	0.390 bc	287 bcd	0.37 de	198.7 a	2237 abc
<i>Fagus</i>	0.474 bc	295 bcd	0.42 cd	125.8 c	1842 cd
<i>Ligustrum</i> 'Argenteum'	0.505 bc	400 a	0.46 cd	160.8 b	1993 bc
<i>Ligustrum</i> 'Aureum'	0.557 b	373 ab	0.47 bc	110.9 cd	2339 ab
<i>Photinia</i> 'Red Robin'	0.805 a	324 abcd	0.56 b	59.6 e	2485 a
<i>Cotoneaster</i>	0.753 a	354 abc	0.64 a	211.9 a	2639 a
LSD (d.f.)	0.1763 (47)	92.1 (47)	0.118 (47)	35.76 (119)	439.6 (50)

728

729

730 Table 4. Mean rainfall volume received within a 40 minute event and volume of  
 731 runoff. Least significant difference (LSD) and degrees of freedom (d.f.) are also  
 732 shown. Different letters next to the means in a column denote statistically significant  
 733 difference between those means ( $P = 0.05$ ).

Treatment	Water volume received (ml) in a 40 min rainfall event	Total runoff volume (ml) after a 40 min rainfall event
Soil	820 a	396 bc
<i>Thuja</i>	3320 b	556 c
<i>Taxus</i>	3890 bc	218 ab
<i>Crataegus</i>	4030 bcd	15 a
<i>Fagus</i>	4660 bcd	187 ab
<i>Ligustrum</i> 'Argenteum'	5100 cd	446 c
<i>Ligustrum</i> 'Aureum'	5170 cd	476 c
<i>Photinia</i> 'Red Robin'	6160 de	638 c
<i>Cotoneaster</i>	7020 e	121 ab
LSD (d.f.)	1296 (39)	376.6 (39)

734

735

736 Table 5. Mean substrate moisture content on days 1, 3, and 5 of the first  
 737 experimental round (22-25 May 2017) along with net CO<sub>2</sub> assimilation and stomatal  
 738 conductance values on day 1 when all plants were well watered. Least significant  
 739 difference (LSD) and degrees of freedom (d.f.) are also shown. Different letters next  
 740 to the means in a column denote statistically significant difference between those  
 741 means (P = 0.05); NS = non-significant.

Treatment	Substrate moisture content (m <sup>3</sup> m <sup>-3</sup> )			Net CO <sub>2</sub> assimilation (μmol m <sup>-2</sup> s <sup>-1</sup> )	Leaf stomatal conductance (mmol m <sup>-2</sup> s <sup>-1</sup> )
	Day 1	Day 3	Day 5	Day 1	Day 1
Bare substrate	0.32	0.27 a	0.23 a	-	-
<i>Cotoneaster</i>	0.26	0.18 b	0.05 b	9.2 a	170.1 a
<i>Crataegus</i>	0.31	0.17 b	0.06 b	6.8 b	103.0 b
<i>Thuja</i>	0.25	0.18 b	0.08 b	5.6 b	94.6 b
LSD (d.f.)	0.068 (47) NS	0.029 (47)	0.019 (47)	1.39 (53)	27.19 (53)

742

743 Table 6. Predicted mean time to runoff (A) and runoff volumes (B) when rainfall was  
 744 applied for 20 min onto troughs where substrate was fully saturated. Data are  
 745 predicted means of three repeated experiments for all treatments and three troughs  
 746 per treatment. Discussion of statistical significance in the body of the text is based on  
 747 Holm p-values.

748 A

Treatment	Predicted mean time to runoff (min)	95% CI: lower bound	95% CI: upper bound
Bare substrate	4.4	-1.4	10.3
<i>Cotoneaster</i>	19.5	14.4	24.6
<i>Crataegus</i>	21.0	15.9	26.1
<i>Thuja</i>	13.2	8.1	18.2

749

750 B

Treatment	Runoff volume at the end of 20 min rainfall (ml)	Runoff volume after 20 min draining (ml)	Runoff volume after 60 min draining (ml)
Bare substrate	256	715	597
<i>Cotoneaster</i>	89	200	97
<i>Crataegus</i>	103	315	118
<i>Thuja</i>	703	779	141

751

752 Table 7. Mean time to runoff (A) and runoff volumes (B) when rainfall was applied for  
 753 60 min onto troughs where substrate was not watered for 3 days. Data are predicted  
 754 means of three repeated experiments for all treatments and three troughs per  
 755 treatment. Discussion of statistical significance in the body of the text is based on  
 756 Holm p-values.

757 A

Treatment	Predicted mean time to runoff (min)	95% CI: lower bound	95% CI: upper bound
Bare substrate	17.8	6.9	28.8
<i>Cotoneaster</i>	31.0	22.9	39.2
<i>Crataegus</i>	38.7	29.4	47.9
<i>Thuja</i>	21.3	12.2	30.5

758

759 B

Treatment	Runoff volume at the end of 60 min rainfall (ml)	Runoff volume after 20 min draining (ml)	Runoff volume after 60 min draining (ml)
Bare substrate	1086	1738	1445
<i>Cotoneaster</i>	1545	471	154
<i>Crataegus</i>	739	255	82
<i>Thuja</i>	2932	943	268

760



761 Figure 1. Setup for the outdoor experiment with model hedges in troughs.

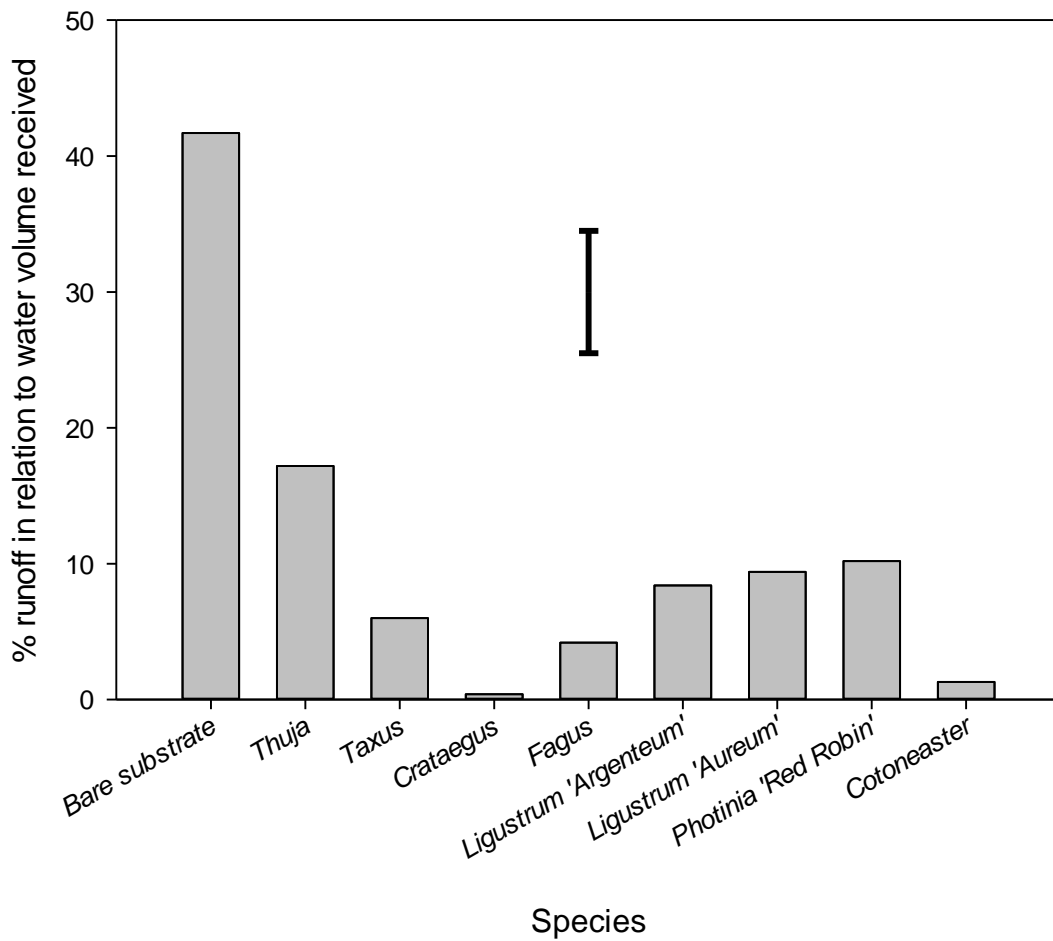


762



763

764 Figure 2. Percent runoff in relation to the rainfall volume received per canopy, after a  
765 40 min simulated rainfall event with the intensity of 28 mm. Rainfall was applied 72 h  
766 after the plants were watered. Values are means of six replicates per plant species  
767 and three replicates for bare soil. Error bar represents least significant difference  
768 between the means (LSD, P = 0.05).



769